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1 INVITED VIEWS IN BASIC AND APPLIED ECOLOGY

2 **Innovations and limits in methods of forecasting global**
3 **environmental change**

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13

14 **Abstract**

15 Environmental science has developed a diverse set of theories, analytical tools and
 16 models to understand and predict ecological responses to human impacts. We review
 17 recent innovations in the family of methods used to forecast global environmental
 18 change, and offer constructive critiques of five common approaches:
 19 phenomenological projections, storyline scenarios, integrated assessment models,
 20 decomposition-identity approaches, and global climate simulations. Overall, there is a
 21 lack of coherent, empirically based validation for many methods and their
 22 assumptions, and only partial incorporation of underlying uncertainties in both
 23 parameter estimates and interrelationships of model components. The greatest
 24 improvements in global environmental forecasting will likely come from a more
 25 systemic approach to quantifying the aggregate socio-economic drivers of the agents
 26 of change, along with better integration of multi-disciplinary approaches.

27

28 **Zusammenfassung**

29 Die Umweltwissenschaft hat vielfältige Theorien, analytische Methoden und Modelle
 30 entwickelt, um ökologische Reaktionen auf anthropogene Einflüsse zu verstehen und
 31 vorherzusagen. Wir untersuchen hier jüngste Innovationen aus der Familie der
 32 Methoden zur Vorhersage von globalen Umweltveränderungen und unterbreiten
 33 konstruktive Kritik zu fünf verbreiteten Forschungsansätzen: phänomenologische
 34 Projektion, "storyline"-Szenarien, integrierte Schätzmodelle, Ansätze zur
 35 Zerlegungsidentität, und Simulationen des globalen Klimas. Insgesamt herrscht ein
 36 Mangel an kohärenter Empirie-gestützter Validierung bei vielen Methoden und ihren
 37 Annahmen. Und die zugrunde liegenden Unsicherheiten, was sowohl.

38 Parameterschätzung als auch Beziehungen zwischen den Modellkomponenten angeht,
39 werden nur teilweise eingearbeitet. Die größten Verbesserungen für globale
40 Umweltvorhersagen werden wahrscheinlich mit einem mehr systemischen Ansatz zur
41 Quantifizierung der aggregierten sozio-öko!
42 nomischen Treiber
43 der bestimmenden Kräfte des Wandels erreicht werden, in Verbindung mit einer
44 engeren Integration von multi-disziplinären Forschungsansätzen. Environmental
45

46 **Keywords:**

47 Projection; Scenario; Integrated Assessment; Decomposition; Multi-criteria Decision
48 Making Analysis; Climate Models; Decoupling.

49

49

50 **Introduction**

51 How might the activities of human civilization drive changes in the Earth system
 52 during the 21st century and beyond? Projections of future environmental states are
 53 inherently constrained by imperfect knowledge and systemic uncertainties in the
 54 drivers of change (Clark et al. 2001). As the famous aphorism goes, all models are
 55 wrong, but some are useful (Box 1979). Forecasts of environmental change are useful
 56 in helping planners trade off the consequences of, and opportunities offered by,
 57 alternative future scenarios (Loftus et al. 2015). Forecasts offer decision makers a way
 58 to anticipate the response of complex systems to chronic stressors or disturbance, and
 59 can permit the evaluation of realistic development pathways to improve conservation
 60 benefit (Ausubel 2000; Leadley et al. 2010; Sala et al. 2000). There are many uses for
 61 scenarios: here we focus primarily on their application to conservation management,
 62 ecology, and their relation to other planning outcomes such economic development. In
 63 this context, the development of ‘what if?’ scenarios can aid in identifying critical
 64 ‘pressure points’ and flexible ‘levers’ for policy, thereby expanding the design space
 65 and opportunities for global conservation while balancing the concessions between
 66 the drive towards equitable human prosperity and the vital need to conserve as much
 67 of our rich natural history and biodiversity as possible.

68 Forecasting should be based on a robust causal framework. One useful
 69 heuristic for conceptualising the linkages between human activities and environmental
 70 transformation is the Driver-Pressure-State-Impact-Response (DPSIR) framework
 71 (Omann et al. 2009). Drivers, including population, consumption, and technology,
 72 determine the aggregate amount of ‘pressures’ (although such a structure lacks
 73 explicit consideration of the role of governance and other aspects of institutional

behaviour that influence the drivers in this framework). Pressures are defined as physical interventions in the environment, and include, for example, land-use change (due to expanding areas of cropland, pasture, biofuels, plantation forests, and built-up land), emissions of greenhouse gases, water extraction, and pollution of air and water (Foley et al. 2005; MEP 2005; Rands et al. 2010). These pressures alter the state of environmental variables (like the distribution of habitats, or the concentration of greenhouse gases in the atmosphere), with attendant impacts on biodiversity (species and populations), in the form of changing abundance, altered geographical distributions, and extinctions (Brook et al. 2008). Responses are the actions taken by humans to address these problems.

Forecasting possible future pathways of biodiversity change (impacts) requires understanding—and modelling—each prior step in this causal chain. Conservation science has developed and validated a rich set of theories and methods to understand and predict the impacts of various human pressures, including population viability analyses, species-area relationships and coupled niche-population models (Botkin et al. 2007; Brook et al. 2000; Ibáñez et al. 2006; Lacy et al. 2013). Conservation science has, however, made less progress on modelling the connections between drivers and pressures. By contrast, in the physical sciences, computer simulations of the Earth System are now routinely used to project emissions of greenhouse gases, the resultant climate change, and its associated risks and impacts (Fordham et al. 2012; Hansen et al. 2007; Lenton et al. 2008). And in the socio-economic realm, integrated assessment models are used to summarize diverse inputs on complex problems such as multi-regional energy projections (Golub et al. 2012; Ostrom 2009).

Despite the progress outlined above, there remains considerable work to do in developing the theoretical and applied tools needed to project and optimize human

development pathways to minimize biodiversity loss from climate change, land-use change, and other pressures. Local interventions like protected areas and payments for ecosystem services can safeguard some of the most valuable elements of biodiversity and ecosystem integrity (Mace et al. 2012). Yet they do little to mitigate the overall level of human pressures, since this is governed primarily by changing patterns of consumption (e.g., demand for material resources) and implementation of new technology (e.g., affecting environmental impacts per unit of production) (Andam et al. 2008; Ausubel 2000; Butchart et al. 2010; Clark et al. 2013). If the hypothesis that technology is a driver (rather than simply a consequence) of social/governance pressures holds true, then the success of biodiversity conservation in the 21st century will depend, to a large extent, on how effectively society can decouple environmental impacts from economic growth and rising human prosperity (Blomqvist et al. 2015; Grubler et al. 1999; UNEP 2011). A failure to achieve this will likely result in an accelerated rate of species extinctions and severe damage to climate and ecosystems—leading to degradation in human health and irreversible loss of natural history (Laurance 2001; Pereira et al. 2010).

Forecasting can play an important role in tackling these problems (some key methods discussed in this paper are outlined in Table 1). To map out future options for managing the planetary environment, it is necessary to incorporate the large uncertainties across both the human dimensions of global change (e.g., technological development, population and demographics, and wealth) (Fig. 1), as well as inherent variability and uncertainty in geophysical and biological processes and feedbacks. The portfolio of past successes and failures in environmental stewardship provides important insights on what *has been achievable*; when integrated with well-structured and parameterized systems models, we then have the critical tools for telling us what

might be possible. Here we explore some of the challenges to projecting change in global-change science.

Phenomenological ('top-down') approaches

Phenomenological models are based on observed relationships between socio-economic and environmental variables (e.g., curves fitted to empirical trend data). These have been used widely across all major impacts of global change, including deforestation, agriculture and pollution (Defries et al. 2010; Ewers et al. 2009; Loh et al. 2005; Sala et al. 2000; Stern et al. 1996; Tilman et al. 2001). For instance, Wright & Muller-Landau (2006) found a strong correlation between rural population density and remaining forest cover across tropical countries and, based on United Nations projections of urbanization and declining rural populations, projected a reduction in pressure on tropical forests in this century. DeFries *et al.* (2010), using a similar methodology, came to the opposite conclusion, finding that urbanization was the socio-economic factor most strongly correlated with forest loss. Tilman *et al.* (2001) used a phenomenological approach to forecast impacts of nitrogen use in agriculture, by extrapolating from historical relationships between nitrogen use, global population, gross domestic product (GDP) and time—estimating that nitrogen use will increase by a factor of 2.7 between 2000 and 2050. The same methodology also underpins a large body of work on the so-called 'Environmental Kuznets Curve', based the proposition that once countries reach a certain income level, environmental impacts peak and then decline (Carson 2009; Jordan 2010). The method used to investigate this question generally involves looking for cross-country statistical relationships between income (represented by GDP) and environmental indicators such as pollution levels or forest loss. Results from these studies are mixed, and often conflicting (Dasgupta et al.

149 2002; Stern 2004).

150 Phenomenological studies like the above have been useful in bringing
 151 attention to socio-economic and technological drivers of environmental change, and
 152 attempting to assess which factors are most influential. Yet, this approach, for a
 153 number of reasons, is strongly limited in its application to forecasting. This is because
 154 it cannot illuminate the mechanisms whereby socio-economic or technological factors
 155 drive environmental change. Its results can therefore be misleading, especially when
 156 extrapolated beyond the historical range of data. For instance, neither Wright &
 157 Muller-Landau (2006) nor DeFries *et al.* (2010) look at the set of interlinked changes
 158 in consumption, production, and trade patterns that are associated with urbanization.
 159 Thus, while urbanization may be correlated with forest loss, phenomenological
 160 studies do not show whether it is causally related, or in which ways. Geist *et al.*
 161 (2002) concluded that these top-down approaches to studying drivers of deforestation
 162 have failed to reveal “any distinct patterns” and thus left the broader question “largely
 163 unanswered”—a conclusion echoed also by DeFries *et al.* (2010). Similarly, the
 164 Tilman *et al.* (2001) extrapolation of global nitrogen use fails to account for regional
 165 patterns in nitrogen use, which tend to follow an inverse U-shaped trend as countries
 166 first adopt synthetic fertilisers and then improve the precision by which it is applied
 167 (Zhang *et al.* 2015). Combining regional trends thus likely yields a plateauing and
 168 even declining trend in nitrogen pollution from agriculture over this century, rather
 169 than a three-fold increase.

170 Studies in the Environmental Kuznets Curve tradition allude to the
 171 mechanisms underpinning improvements in environmental quality in qualitative
 172 terms, but do not analyse them directly. Thus the method does not differentiate
 173 between technological improvements *per se*, and displacement of environmentally

harmful activities abroad (Ansuategi & Perrings 2000). It also overlooks the often significant differences in environmental pressures between countries at similar income levels, which seem to have resulted from path-dependent economic and technological choices rather than differences in economic growth.

‘Storyline’ scenarios

Storyline scenarios have been used extensively by the Intergovernmental Panel on Climate Change in their five Assessment Reports, and underpinned the ‘Scenarios’ volume of the 2005 Millennium Ecosystem Assessment (MA; MEP 2005), the Global Biodiversity Outlook (CBD 2013), and many other assessments and horizon scans. Indeed, this approach has become the main analytical lens through which the future of global biodiversity and ecosystems has been perceived and interpreted.

Storyline scenarios start with a narrative that defines a hypothetical pathway for population growth and economic development, as well as technological and institutional change. In the case of the MA, the scenarios are framed along two axes: degree of globalisation and proactive versus reactive policies—yielding four different storylines (MEP 2005). These assumptions then serve as input to complex cross-disciplinary simulations—in most cases a form of ‘bottom-up’ economic analysis called Integrated Assessment Modelling (IAM, see next section)—which can be used to project (i) the magnitude of pressures like land-use change or pollution, and (ii) resultant changes in biodiversity and ecosystem integrity.

Storyline approaches, although intellectually appealing and easy to communicate, almost certainly underestimate the range of plausible future outcomes (Leadley et al. 2010) and typically say little about the feasibility of implementation (Loftus et al. 2015). For example, projections for increases in global agricultural area

199 fall within a relatively narrow 11% range for all Millennium Ecosystem Assessment
 200 scenarios. This seems to be due to compensatory mechanisms whereby inputs that
 201 lead to increased land use (e.g., vastly expanded use of crops for bioenergy) are
 202 combined in the same scenario with other parameters that reduce land use (e.g.,
 203 reduced meat consumption and higher agricultural yields). Similarities across
 204 ‘different’ storyline scenarios are exacerbated further by use of the same IAMs for
 205 estimating drivers and biodiversity responses (Tallis & Kareiva 2006). Furthermore, a
 206 well-established psychological effect exists whereby a high level of detail, such as
 207 exists for any of the MA storylines, leads to a high level of perceived likelihood of the
 208 scenario coming true (Morgan & Keith 2008). Thus, contrary to the stated objective of
 209 typical storyline scenarios, this method might often lead to constrained thinking
 210 around different options and pathways. Perhaps most critically, the fact that storyline
 211 scenarios come as a fixed bundle of parameters also makes it nearly impossible to
 212 gauge the effects or sensitivity of the environmental outcomes to individual policy
 213 options, such as organic versus conventional farming, or wind power versus biomass.

214

215 **Integrated Assessment Models**

216 Integrated Assessment Models are closely linked to storylines in that they often base
 217 their projections on assumptions about drivers like population, GDP, and technology
 218 derived from storylines, IAMs leverage well-verified economic approaches such as
 219 computable general equilibrium models to assimilate data on how individual
 220 economies might respond to changes in policy, technology, or cross-border factors
 221 (Fig. 2), and then aggregate these results to produce plausible bottom-up scenarios of
 222 change (Garnaut 2008; Valin et al. 2013). This is typically achieved using recursive-
 223 dynamic approaches, based on mechanistic relationships, which are solved

224 sequentially. These models can also be used for probabilistic assessments of policy,
225 especially in situations where uncertainty is accepted to be high (such as for
226 evaluating interventions to mitigate climate change; Mastrandrea & Schneider 2004).
227 The philosophy of IAMs is relatively blind to disciplinary borders and typically
228 involves inputs from a diversity of specialized experts. Widely used examples in the
229 climate-energy policy realm include MiniCAM, MERGE and IGSM (Clarke et al.
230 2007).

231 Although IAM results provide cohesive information that can assist policy
232 makers in developing more transparent approaches to scenario analysis, they have the
233 disadvantage of being (by definition) quite complex, heavily assumption driven, and
234 can be rather opaque (Pielke et al. 2008; van der Sluijs 2002). For instance, modelling
235 the stabilization pathways for greenhouse-gas emissions involves three broad items: a
236 reduction in end-use demand (efficiency and conservation), an increase in carbon-free
237 energy to replace fossil fuels (e.g., renewables and nuclear), and some switch-over of
238 fossil fuels to carbon capture and storage (CCS) (Hoffert et al. 2002). On this basis,
239 the IAMs attempt to resolve cost-optimized scenarios that meet defined emissions
240 targets, usually in decadal bands through to mid- or end-of century (Clarke et al.
241 2007; Wise et al. 2009).

242 The principal challenge in projecting something like greenhouse gas emissions
243 using IAMs is to realistically characterize both socio-political choices (e.g. when and
244 at what level a carbon price or low-carbon-energy production credit is implemented,
245 community antagonism against widespread use of nuclear fission or building of wind
246 farms) and the scientific-economic evolution of, and deployment rates for, the
247 underlying technologies themselves (e.g., engineering efficiencies of energy
248 conversion, dispatchability of the resource for load balancing, or cost-reduction

249 curves for grid-scale renewables with integrated storage) (Lenzen et al. 2013; Utgikar
 250 & Scott 2006). This is important, because these uncertainties and assumptions are not
 251 only difficult to constrain *a priori*, they also cascade into a wide range of possible
 252 climate-forcing scenarios (which are fed into global climate models; GCMs) (Wigley
 253 et al. 2009). As a consequence, methods that build upon the intrinsic uncertainties in
 254 the GCMs typically result in (necessarily) wide bounds of probability for projections
 255 of habitat change and species distributions when forecasting biodiversity responses,
 256 thus appropriately reflecting our high degree of uncertainty about many future
 257 ecological outcomes (Botkin et al. 2007; Fordham et al. 2011).

258

259 **Decomposition and Identity approaches**

260 The alternative to the phenomenological and storyline approaches is to apply a suite
 261 of relatively simple, bottom-up decompositions of human drivers into a set of
 262 multiplicative factors, using a set of methods associated with ecological economics
 263 and industrial ecology (Duchin & Lange 1995; Steinberger et al. 2010; Thomas et al.
 264 2003; Wiedmann 2009). This approach seeks to make all assumptions and exogenous
 265 inputs into the models transparent. Drawing on the classical IPAT formula (Impact =
 266 Population x Affluence x Technology) (Chertow 2001; Ehrlich & Holdren 1971),
 267 Waggoner & Ausubel (2002) developed a mathematical identity, ImPACT (with C
 268 being consumer use per GDP), wherein environmental impacts are the product of
 269 population, income, intensity of use (material throughput per unit of income), and
 270 intensity of impact (environmental impact per unit material throughput). This type of
 271 ‘decomposition’ (i.e., breakdown of general models into more fined-grained factors)
 272 (Ang 2004) has been applied extensively to the study of energy and greenhouse-gas
 273 emissions, under the umbrella of the Kaya Identity, where total emissions are a

274 product of population, income, energy intensity (energy use per unit income), and
 275 emissions intensity (emissions per unit energy) (Hamilton & Turton 2002; Rosa &
 276 Dietz 2012). The framework and precise factors used are flexible, provided they form
 277 an identity; for instance transport-sector emissions can be decomposed into passenger-
 278 km, transport modes, carbon and energy intensity, and fuel mix (Stern 1997).

279 The idea behind this approach to projecting change is that demand forecasts
 280 for key economic goods, as outlined above, should be combined with a rigorous
 281 analysis of technological trajectories and options to estimate aggregate environmental
 282 impacts. The benefit of the decomposition-identity approach is that the contribution of
 283 each factor to the aggregate change in impacts can be determined readily, with general
 284 models broken down into increasingly fined-grained factors, thereby allowing direct
 285 investigation of the sensitivity of outcomes to different policy levers. The method has
 286 also served to highlight how a combination of declining intensity of use
 287 (dematerialization) and intensity of impact (i.e., increasing technical efficiency) can
 288 offset some or all of the pressure from growing population and economic activity,
 289 thereby decoupling environmental impacts like land use and water consumption from
 290 economic growth (Ausubel & Waggoner 2008; Ausubel et al. 2012; Voet et al. 2005).

291 However, as York *et al.* (2003) have pointed out, rudimentary mathematical
 292 identities like ImPACT, while useful accounting tools, have limited utility in
 293 forecasting. Although it encourages mechanistic ‘bottom-up’ approaches to
 294 forecasting, the aggregated parameters have to be assumed, rather than being data-
 295 driven; interactions between factors are not accounted for, and growth functions are
 296 typically assumed to be exponential. Moreover, this method has primarily been
 297 applied to very high levels of aggregation, often global, thereby omitting many lower-
 298 scale patterns and dynamics. The STIRPAT (Stochastic Impacts by Regression on

Population, Affluence and Technology) method is a step forward, because it allows for data-driven fitting of coefficients and sensitivity evaluation (Liddle & Lung 2010). However, it does not offer a fully adequate and comprehensive method, since, for instance, it ignores model selection and does not make use of prior information. For more accurate forecasting, the technology factor must be disaggregated into distinct processes or transformations, each with their own theoretical limits, learning curves, and variation across systems and countries. Technological change has a second component in addition to incremental improvement: understanding the benefits and limits of *substitution*, whereby one technology replaces another (Chang & Baek 2010; Grubler et al. 1999; Mace 2012).

Global Climate and General Ecosystem Models

Predicting future impacts of climate change on biodiversity illustrates the many challenges involved in forecasting the interlinked components of the causal chain, from drivers like consumption and technology, to pressures (greenhouse gas emissions), to changes in the state of the global climate system, and finally to impacts on biodiversity. Indeed, in seeking to bracket the range of plausible anthropogenically forced scenarios, climate modellers typically employ a combination of mechanistic and scenario-based approaches to projecting change (Moss et al. 2010) (Fig. 1). They assess the skill of global climate models (GCMs) based on validation against historical data (Fordham et al. 2013). The linking of GCM outputs to forecasts of biodiversity response necessitates estimates of both mean trends in climatic variables like temperature and precipitation, and also a characterization of their variability, extremes, and key uncertainties in the underpinning models (Botkin et al. 2007; Brook et al. 2009).

324 One of the challenges in projecting climate change lies in the structural
325 adequacy and spatial resolution of the atmosphere-ocean global circulation models
326 that underpin the simulations (Wigley & Raper 2001). This stems from modellers'
327 incomplete understanding (and weak parameterization) of crucial mechanisms such as
328 heat transport in the ocean, cloud formation, and boundary-layer formulations (IPCC
329 2013). Another source of ambiguity is in how well-known geophysical processes and
330 less-certain amplifying or diminishing feedbacks should be best represented and
331 integrated, which results in a band of nearly irreducible uncertainty in the equilibrium
332 climate sensitivity of different GCMs (Hansen et al. 2007). A reassuring result of the
333 last few decades of work in this area has been steady improvements in both the short-
334 term forecasting (used for weather predictions) and longer-term hindcasting ability of
335 GCMs, thanks to greatly increased spatial resolution and inclusion of increasingly
336 complex features (e.g., layered-ocean modelling, carbon-cycle processes, and explicit
337 incorporation of dynamic vegetation and ice-sheet models) (Reichler & Kim 2008).
338 These enhancements have been made possible by the exponential recent growth in
339 computer power, and should continue for many years.

340 Even accepting that current GCMs will remain an imperfect simplification of
341 the highly complex Earth system for years to come, we can still make progress in the
342 challenge of more objectively representing future change. A well-regarded method is
343 to accept that there are a range of potentially valid ways of simulating these complex
344 systems and so treat the diversity of approaches tried by different climate-modelling
345 communities as an advantage, by pooling their probabilistic GCM results in an
346 'ensemble' forecast (Tebaldi & Knutti 2007). This combining of multi-model output
347 can include the assignment of differential weightings to alternative models on the
348 basis of, say, their 'skill' score with respect to their ability to simulate past climates

349 (Gleckler et al. 2008). Besides global metrics, this skill ranking can also be
350 disaggregated at regional scales and separately for different outputs (e.g., some
351 models seem to be better at reconstructing changes in temperature, whereas others are
352 superior at reconstructing past interannual variability in precipitation (Scherrer 2011),
353 even though their temperature forecasts may be sub-par). Recent advances in user-
354 friendly emulation software (e.g. MAGICC/SCENGEN and GridMapper) have more
355 readily opened the application of the climate-ensembling approach to ecologically
356 focused end-users (Fordham et al. 2012) (Fig. 3). Another simpler but related
357 approach relies on projecting change using both the best-performing and the most
358 extreme models (for a given output), to attempt to encompass the full range of
359 possible futures using selected inter-model comparisons.

360 An additional component of uncertainty in climate models is in characterizing
361 the likely future pathways of climate forcing factors, which includes long-lived
362 greenhouse gases such as carbon dioxide and methane, aerosol loads, the capacity of
363 the oceans, vegetation and soil to continue to act as a net carbon sink, as well as the
364 dynamics of natural variability in ocean circulation, volcanoes, and solar output
365 (Wigley et al. 2009). This can be done by assuming little or no long-term trend in
366 volcanic or solar forcing, treating observed regional fluctuations such as El Niño
367 Southern Oscillation (ENSO) as canonical or emergent properties, and exploring the
368 climatic implications (over the next few centuries, and for the stabilized equilibrium
369 condition) of a range of different ‘storylines’ of future global energy and emissions
370 profiles (from business-as-usual to explicit mitigation policies). Forecasts then can be
371 expressed either via socio-economic pathways using IAMs (Nakicenovic & Swart
372 2000) or selected from a large suite of possible scenarios on the basis of their resultant
373 radiative forcing potential (e.g. the ‘representative concentration pathways’ of the

374 Intergovernmental Fifth Assessment Report; IPCC 2013).

375 It is obvious from the above discussion that the development of General
376 Circulation Models for climate simulation has advanced considerably over the last
377 few decades, and these arguably offer salutary lessons for the design of analogous
378 system-level biodiversity-response models. For instance, one promising recent
379 approach is the ‘General Ecosystem Model’ (GEM), developed in a collaboration
380 between the United Nations Environment Program, World Conservation Monitoring
381 Centre, and Microsoft Research. The ambitious goal of this global simulation model is
382 to capture the fundamental ecological processes that affect all life on Earth as a
383 ‘virtual biosphere’ using an interactive mathematical simulation (called the Madingley
384 Model: madingleymodel.org). The code has been released as open source, and is
385 undergoing testing, validation and ongoing community development (Harfoot et al.
386 2014). An ongoing challenge for such GEMs will be solving the challenges of
387 integrating human decision-making processes and including institutional complexities
388 into the underpinning regional- and global-level processes (Geographical Sciences
389 Committee 2014; Rounsevell et al. 2013).

390

391 **Conclusions**

392 A range of useful methods has been developed to project global change. Yet, as
393 reviewed above, there are clearly limitations with all of these lines of attack. Perhaps
394 most pressingly, global-change science still lacks a coherent, empirically based,
395 statistically robust, and transparent methodology to understand and forecast human
396 drivers of land-use change (and associated impacts) and in turn connect this to
397 biodiversity responses at regional to global scales. This constrains our understanding
398 of both the long-term prospects of biodiversity change and on what interventions

399 might be most effective. At higher levels of aggregation, patterns in consumption and
400 use of technology over time and between countries and regions constitute perhaps the
401 most readily identifiable and consistent bases for projecting change.

402 To increase confidence in our representations of the future, we must seek
403 broad expert elicitation (for proper representation of different disciplinary
404 perspectives) and ensure that models (and assumptions) are validated against robust
405 historical data on key uncertainties, such as rates of technology uptake and barriers to
406 deployment. Confidence in the likelihood of scenarios can be enhanced by analysis of
407 the short-term impact of already announced government policy targets (assuming they
408 are implemented in full, e.g., IEA 2010) or by reference to the envisaged goals from
409 organizations or businesses with a strong track record at delivery (Chang & Baek
410 2010; Nicholson et al. 2011; Smil 2010). Quantitative tools like multi-criteria
411 decision-making analysis, decomposition and input-output models (Hong et al. 2013;
412 Rose & Casler 1996) offer a particularly useful pathway for ensuring high levels of
413 robustness and openness in such validation. Models should also be tested repeatedly
414 against real-world data on patterns and trends—just like hypotheses—to learn from
415 their failures as much as their successes (Brook et al. 2002; Grimm et al. 2005).
416 Crucially, the modelling of aggregate drivers provides boundary conditions for more
417 local contexts, which are often more complex, and so can complement and support
418 studies and methodologies at lower spatial scales. To further improve our forecasting,
419 mechanistic approaches based on robust data—on demographics, incomes, industrial
420 sectors, per-capita consumption of key resources, trade, land use, technical
421 efficiencies of production methods, pollution, and so on—will need to come from
422 many sources: global to national reporting inventories, remote sensing, and biological
423 surveys, among others. These sources should be set up in a way that is readily

424 interrogated with relational databasing.

425 A transformation is underway in research on global-change science, driven by
 426 ready access to ‘big data’ from observational and experimental networks, ongoing
 427 growth in computational power, and complementary advances in statistical and
 428 optimisation methodologies. What is critically needed to complement these
 429 developments are validated, mechanistic models of the drivers of global change,
 430 integrated with approaches that are flexible enough to capture key uncertainties and
 431 complex interrelationships, but simple and transparent enough to be applied
 432 efficiently for optimising decision-making and testing the sensitivity of assumptions.

433

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437

438 **References**

- 439 Andam, K.S., Ferraro, P.J., Pfaff, A., Sanchez-Azofeifa, G.A., & Robalino, J.a.
 440 (2008). Measuring the effectiveness of protected area networks in reducing
 441 deforestation. *Proceedings of the National Academy of Sciences of the United States*
 442 *of America*, 105, 16089-16094.
- 443 Ang, B.W. (2004). Decomposition analysis for policymaking in energy. *Energy*
 444 *Policy*, 32, 1131-1139.
- 445 Ansuategi, A., & Perrings, C. (2000). Transboundary externalities in the
 446 environmental transition hypothesis. *Environmental and Resource Economics*, 17,
 447 353-373.
- 448 Ausubel, J.H. (2000). The great reversal: nature's chance to restore land and sea.
 449 *Technology in Society*, 22, 289-301.

- 450 Ausubel, J.H., & Waggoner, P.E. (2008). Dematerialization: Variety, caution, and
451 persistence. *Proceedings of the National Academy of Sciences of the United States of*
452 *America*, 105, 12774-12779.
- 453 Ausubel, J.H., Wernick, I.K., & Waggoner, P.E. (2012). Peak farmland and the
454 prospect for land sparing. *Population and Development Review*, 38, 221-242.
- 455 Blomqvist, L., Nordhaus, T., & Shellenberger, M. (2015). *Nature Unbound:*
456 *Decoupling for Conservation*. Oakland, CA: Breakthrough Institute.
- 457 Botkin, D.B., Saxe, H., Araujo, M.B., Betts, R., Bradshaw, R.H.W., Cedhagen, T.,
458 Chesson, P., Dawson, T.P., Etterson, J.R., Faith, D.P., Ferrier, S., Guisan, A., Hansen,
459 A.S., Hilbert, D.W., Loehle, C., Margules, C., New, M., Sobel, M.J., & Stockwell,
460 D.R.B. (2007). Forecasting the effects of global warming on biodiversity. *BioScience*,
461 57, 227-236.
- 462 Box, G.E. (1979). Robustness in the strategy of scientific model building. *Robustness*
463 *in statistics* (pp. 201-236): Academic Press.
- 464 Brook, B.W., Akcakaya, H.R., Keith, D.A., Mace, G.M., Pearson, R.G., & Araujo,
465 M.B. (2009). Integrating bioclimate with population models to improve forecasts of
466 species extinctions under climate change. *Biology Letters*, 5, 723-725.
- 467 Brook, B.W., Burgman, M.A., Akcakaya, H.R., O'grady, J.J., & Frankham, R. (2002).
468 Critiques of PVA ask the wrong questions: throwing the heuristic baby out with the
469 numerical bath water. *Conservation Biology*, 16, 262-263.
- 470 Brook, B.W., O'Grady, J.J., Chapman, A.P., Burgman, M.A., Akçakaya, H.R., &
471 Frankham, R. (2000). Predictive accuracy of population viability analysis in
472 conservation biology. *Nature*, 404, 385-387.
- 473 Brook, B.W., Sodhi, N.S., & Bradshaw, C.J.A. (2008). Synergies among extinction
474 drivers under global change. *Trends in Ecology & Evolution*, 23, 453-460.
- 475 Butchart, S.H.M., Walpole, M., Collen, B., van Strien, A., Scharlemann, J.P.W.,
476 Almond, R.E.A., Baillie, J.E.M., Bomhard, B., Brown, C., Bruno, J., Carpenter, K.E.,
477 Carr, G.M., Chanson, J., Chenery, A.M., Csirke, J., Davidson, N.C., Dentener, F.,

- 478 Foster, M., Galli, A., Galloway, J.N., Genovesi, P., Gregory, R.D., Hockings, M.,
479 Kapos, V., Lamarque, J.-F., Leverington, F., Loh, J., McGeoch, M.A., McRae, L.,
480 Minasyan, A., Hernández Morcillo, M., Oldfield, T.E.E., Pauly, D., Quader, S.,
481 Revenga, C., Sauer, J.R., Skolnik, B., Spear, D., Stanwell-Smith, D., Stuart, S.N.,
482 Symes, A., Tierney, M., Tyrrell, T.D., Vié, J.-C., & Watson, R. (2010). Global
483 biodiversity: indicators of recent declines. *Science*, 328, 1164-1168.
- 484 Carson, R.T. (2009). The Environmental Kuznets Curve: Seeking empirical regularity
485 and theoretical structure. *Review of Environmental Economics and Policy*, 4, 3-23.
- 486 CBD. (2013). *Convention on Biological Diversity: Global Biodiversity Outlook 3*.
487 <http://www.cbd.int/gbo3/>.
- 488 Chang, Y.S., & Baek, S.J. (2010). Limit to improvement: Myth or reality? Empirical
489 analysis of historical improvement on three technologies influential in the evolution of
490 civilization. *Technological Forecasting and Social Change*, 77, 712-729.
- 491 Chertow, M.R. (2001). The IPAT Equation and Its Variants and Environmental
492 Impact. *Journal of Industrial Ecology*, 4, 13-29.
- 493 Clark, J.S., Carpenter, S.R., Barber, M., Collins, S., Dobson, A., Foley, J.a., Lodge,
494 D.M., Pascual, M., Pielke, R., Pizer, W., Pringle, C., Reid, W.V., Rose, K.a., Sala, O.,
495 Schlesinger, W.H., Wall, D.H., & Wear, D. (2001). Ecological forecasts: an emerging
496 imperative. *Science*, 293, 657-660.
- 497 Clark, N.E., Boakes, E.H., McGowan, P.J.K., Mace, G.M., & Fuller, R.A. (2013).
498 Protected areas in South Asia have not prevented habitat loss: a study using historical
499 models of land-use change. *PLoS ONE*, 8, e65298.
- 500 Clarke, L., Edmonds, J., Jacoby, J., Pitcher, H., Reilly, J., Richels, R., Parson, E.,
501 Burkett, V., Fisher-Vanden, K., Keith, D., Mearns, L., Rosenzweig, C., & Webster,
502 M. (2007). *CCSP Scenarios of Greenhouse Gas Emissions and Atmospheric*
503 *Concentrations*. Washington D.C.: Department of Energy, Office of Biological &
504 Environmental Research.
- 505 Dasgupta, S., Laplante, B., Wang, H., & Wheeler, D. (2002). Confronting the
506 Environmental Kuznets Curve. *Journal of Economic Perspectives*, 16, 147-168.

- 507 Defries, R.S., Rudel, T., Uriarte, M., & Hansen, M. (2010). Deforestation driven by
508 urban population growth and agricultural trade in the twenty-first century. *Nature*
509 *Geoscience*, 3, 178-181.
- 510 Duchin, F., & Lange, G.-M. (1995). The Future of the Environment: Ecological
511 Economics & Technological Change. 222.
- 512 Ehrlich, P.R., & Holdren, J.P. (1971). Impacts of population growth. *Science*, 171,
513 1212-1217.
- 514 Ewers, R.M., Scharlemann, J.P.W., Balmford, A., & Green, R.E. (2009). Do increases
515 in agricultural yield spare land for nature? *Global Change Biology*, 15, 1716-1726.
- 516 Foley, J.a., DeFries, R.S., Asner, G.P., Barford, C., Bonan, G., Carpenter, S.R.,
517 Chapin, F.S., Coe, M.T., Daily, G.C., Gibbs, H.K., Helkowski, J.H., Holloway, T.,
518 Howard, E.a., Kucharik, C.J., Monfreda, C., Patz, J.a., Prentice, I.C., Ramankutty, N.,
519 & Snyder, P.K. (2005). Global consequences of land use. *Science*, 309, 570-574.
- 520 Fordham, D.A., Akçakaya, H.R., Araújo, M.B., Keith, D.A., & Brook, B.W. (2013).
521 Tools for integrating range change, extinction risk and climate change information
522 into conservation management. *Ecography*, 36, 956-964.
- 523 Fordham, D.A., Wigley, T.M.L., & Brook, B.W. (2011). Multi-model climate
524 projections for biodiversity risk assessments. *Ecological Applications*, 21, 3317-3331.
- 525 Fordham, D.A., Wigley, T.M.L., & Brook, B.W. (2012). Strengthening forecasts of
526 climate change impacts with multi-model ensemble averaged projections using
527 MAGICC/SCENGEN 5.3. *Ecography*, 35, 4-8.
- 528 Garnaut, R. (2008). *The Garnaut Climate Change Review*. Cambridge University
529 Press, Melbourne.
- 530 Geist, H.J., Lambin, E.F., Forests, R., Disappearing, A.R.E., The, A.S., Of, R.,
531 Pressures, M., Local, B., In, A., Combinations, V., Different, I.N., & Locations, G.
532 (2002). Proximate causes and underlying driving forces of tropical deforestation.
533 *BioScience*, 52, 143-150.

- 534 Geographical Sciences Committee. (2014). *Advancing Land Change Modeling:*
535 *Opportunities and Research Requirements*. National Academies Press.
- 536 Gleckler, P.J., Taylor, K.E., & Doutriaux, C. (2008). Performance metrics for climate
537 models. *Journal of Geophysical Research*, 113, D06104.
- 538 Golub, A.a., Henderson, B.B., Hertel, T.W., Gerber, P.J., Rose, S.K., & Sohngen, B.
539 (2012). Global climate policy impacts on livestock, land use, livelihoods, and food
540 security. *Proceedings of the National Academy of Sciences of the United States of*
541 *America*, 110, 20894-20899.
- 542 Grimm, V., Revilla, E., Berger, U., Jeltsch, F., Mooij, W.M., Railsback, S.F., Thulke,
543 H.-H., Weiner, J., Wiegand, T., & DeAngelis, D.L. (2005). Pattern-oriented modeling
544 of agent-based complex systems: lessons from ecology. *Science*, 310, 987-991.
- 545 Grubler, A., Nakicenovic, N., & Victor, D.G. (1999). Modeling technological change:
546 implications for the global environment. *Annual Review of Energy and the*
547 *Environment*, 24, 545-569.
- 548 Hamilton, C., & Turton, H. (2002). Determinants of emissions growth in OECD
549 countries. *Energy Policy*, 30, 63-71.
- 550 Hansen, J., Sato, M., Kharecha, P., Russell, G., Lea, D.W., & Siddall, M. (2007).
551 Climate change and trace gases. *Philosophical Transactions of the Royal Society of*
552 *London. Series A, Mathematical Physical and Engineering Sciences*, 365, 1925-1954.
- 553 Harfoot, M.B., Newbold, T., Tittensor, D.P., Emmott, S., Hutton, J., Lyutsarev, V.,
554 Smith, M.J., Scharlemann, J.P., & Purves, D.W. (2014). Emergent global patterns of
555 ecosystem structure and function from a mechanistic general ecosystem model. *PLoS*
556 *Biology*, 12, e1001841.
- 557 Hoffert, M.I., Caldeira, K., Benford, G., Criswell, D.R., Green, C., Herzog, H., Jain,
558 A.K., Kheshgi, H.S., Lackner, K.S., Lewis, J.S., Lightfoot, H.D., Manheimer, W.,
559 Mankins, J.C., Mauel, M.E., Perkins, L.J., Schlesinger, M.E., Volk, T., & Wigley,
560 T.M.L. (2002). Advanced technology paths to global climate stability: Energy for a
561 greenhouse planet. *Science*, 298, 981-987.

- 562 Hong, S., Bradshaw, C.J.A., & Brook, B.W. (2013). Evaluating options for the future
563 energy mix of Japan after the Fukushima nuclear crisis. *Energy Policy*, 56, 418-424.
- 564 Ibáñez, I., Clark, J.S., Dietze, M.C., Feeley, K., Hersh, M., LaDeau, S., McBride, A.,
565 Welch, N.E., & Wolosin, M.S. (2006). Predicting biodiversity change: outside the
566 climate envelope, beyond the species-area curve. *Ecology*, 87, 1896-1906.
- 567 IEA. (2010). *World Energy Outlook 2010*. <http://www.worldenergyoutlook.org/>.
- 568 IPCC. (2013). *Intergovernmental Panel on Climate Change - Fifth Assessment Report*
569 (AR5). Geneva, Switzerland, <http://www.ipcc.ch>.
- 570 Jordan, B.R. (2010). The Environmental Kuznets Curve: Preliminary Meta-Analysis
571 of Published Studies , 1995-2010.
- 572 Lacy, R.C., Miller, P.S., Nyhus, P.J., Pollak, J.P., Raboy, B.E., & Zeigler, S.L.
573 (2013). Metamodels for Transdisciplinary Analysis of Wildlife Population Dynamics.
574 *PLoS ONE*, 8, e84211.
- 575 Laurance, W.F. (2001). Future shock: forecasting a grim fate for the Earth. *Trends in*
576 *Ecology & Evolution*, 16, 531-533.
- 577 Leadley, P., Pereira, H., & Alkemade, R. (2010). *Biodiversity Scenarios: Projections*
578 *of 21st Century Change in Biodiversity and Associated Ecosystem Services*. Montreal:
579 Secretariat of the Convention on Biological Diversity.
- 580 Lenton, T.M., Held, H., Kriegler, E., Hall, J.W., Lucht, W., Rahmstorf, S., &
581 Schellnhuber, H.J. (2008). Tipping elements in the Earth's climate system.
582 *Proceedings of the National Academy of Sciences of the United States of America*,
583 105, 1786-1793.
- 584 Lenzen, M., Dey, C., Foran, B., Widmer-Cooper, A., Ohlemüller, R., Williams, M., &
585 Wiedmann, T. (2013). Modelling interactions between economic activity, greenhouse
586 gas emissions, biodiversity and agricultural production. *Environmental Modeling &*
587 *Assessment*, 18, 377-416.
- 588 Liddle, B., & Lung, S. (2010). Age-structure, urbanization, and climate change in
589 developed countries: revisiting STIRPAT for disaggregated population and

- 590 consumption-related environmental impacts. *Population and Environment*, 31, 317-
591 343.
- 592 Loftus, P.J., Cohen, A.M., Long, J.C.S., & Jenkins, J.D. (2015). A critical review of
593 global decarbonization scenarios: what do they tell us about feasibility? *Wiley*
594 *Interdisciplinary Reviews: Climate Change*, 6, 93-112.
- 595 Loh, J., Green, R.E., Ricketts, T., Lamoreux, J., Jenkins, M., Kapos, V., & Randers, J.
596 (2005). The Living Planet Index: using species population time series to track trends
597 in biodiversity. *Philosophical Transactions of the Royal Society of London. Series B*,
598 *Biological sciences*, 360, 289-295.
- 599 Mace, G.M. (2012). The limits to sustainability science: ecological constraints or
600 endless innovation? *PLoS Biology*, 10, e1001343.
- 601 Mace, G.M., Norris, K., & Fitter, A.H. (2012). Biodiversity and ecosystem services: a
602 multilayered relationship. *Trends in Ecology & Evolution*, 27, 19-26.
- 603 Mastrandrea, M.D., & Schneider, S.H. (2004). Probabilistic integrated assessment of
604 "dangerous" climate change. *Science*, 304, 571-575.
- 605 MEP. (2005). *Millenium Ecosystem Assessment: Ecosystems and Human Well-being:*
606 *Scenarios*. Washington, DC: Island Press.
- 607 Morgan, M.G., & Keith, D.W. (2008). Improving the way we think about projecting
608 future energy use and emissions of carbon dioxide. *Climatic Change*, 90, 189-215.
- 609 Moss, R.H., Edmonds, J.A., Hibbard, K.A., Manning, M.R., Rose, S.K., van Vuuren,
610 D.P., Carter, T.R., Emori, S., Kainuma, M., Kram, T., Meehl, G.A., Mitchell, J.F.B.,
611 Nakicenovic, N., Riahi, K., Smith, S.J., Stouffer, R.J., Thomson, A.M., Weyant, J.P.,
612 & Wilbanks, T.J. (2010). The next generation of scenarios for climate change research
613 and assessment. *Nature*, 463, 747-756.
- 614 Nakicenovic, N., & Swart, R. (Eds.). (2000). *Special Report on Emissions Scenarios*.
615 Cambridge University Press, UK.

- 616 Nicholson, M., Biegler, T., & Brook, B.W. (2011). How carbon pricing changes the
617 relative competitiveness of low-carbon baseload generating technologies. *Energy*, 36,
618 305-313.
- 619 Omann, I., Stocker, A., & Jäger, J. (2009). Climate change as a threat to biodiversity:
620 An application of the DPSIR approach. *Ecological Economics*, 69, 24-31.
- 621 Ostrom, E. (2009). A general framework for analyzing sustainability of social-
622 ecological systems. *Science*, 325, 419-422.
- 623 Pereira, H.M., Leadley, P.W., Proença, V., Alkemade, R., Scharlemann, J.P.W.,
624 Fernandez-Manjarrés, J.F., Araújo, M.B., Balvanera, P., Biggs, R., Cheung, W.W.L.,
625 Chini, L., Cooper, H.D., Gilman, E.L., Guénette, S., Hurtt, G.C., Huntington, H.P.,
626 Mace, G.M., Oberdorff, T., Revenga, C., Rodrigues, P., Scholes, R.J., Sumaila, U.R.,
627 & Walpole, M. (2010). Scenarios for global biodiversity in the 21st century. *Science*,
628 330, 1496-1501.
- 629 Pielke, R., Wigley, T., & Green, C. (2008). Dangerous assumptions. *Nature*, 452,
630 531-532.
- 631 Rands, M.R.W., Adams, W.M., Bennun, L., Butchart, S.H.M., Clements, A., Coomes,
632 D., Entwistle, A., Hodge, I., Kapos, V., Scharlemann, J.P.W., Sutherland, W.J., &
633 Vira, B. (2010). Biodiversity conservation: challenges beyond 2010. *Science*, 329,
634 1298-1303.
- 635 Reichler, T., & Kim, J. (2008). How well do coupled models simulate today's
636 climate? *Bulletin of the American Meteorological Society*, 89, 303-311.
- 637 Rosa, E.A., & Dietz, T. (2012). Human drivers of national greenhouse-gas emissions.
638 *Nature Climate Change*, 2, 581-586.
- 639 Rose, A., & Casler, S. (1996). Input-output structural decomposition analysis: a
640 critical appraisal. *Economic Systems Research*, 8, 33-62.
- 641 Rounsevell, M., Arneth, A., Brown, D., de Noblet-Ducoudré, N., Ellis, E., Finnigan,
642 J., Galvin, K., Grigg, N., Harman, I., & Lennox, J. (2013). Incorporating human

- behaviour and decision making processes in land use and climate system models. GLP Report.
- Sala, O.E., Chapin, F.S., Armesto, J.J., Berlow, E., Bloomfield, J., Dirzo, R., Huber-Sanwald, E., Huenneke, L.F., Jackson, R.B., Kinzig, a., Leemans, R., Lodge, D.M., Mooney, H.a., Oesterheld, M., Poff, N.L., Sykes, M.T., Walker, B.H., Walker, M., & Wall, D.H. (2000). Global biodiversity scenarios for the year 2100. *Science*, 287, 1770-1774.
- Scherrer, S.C. (2011). Present-day interannual variability of surface climate in CMIP3 models and its relation to future warming, I. *International Journal of Climatology*, 31, 1518-1529.
- Smil, V. (2010). *Energy Myths and Realities: Bringing Science to the Energy Policy Debate*. Washington, D.C.: AEI Press.
- Steinberger, J.K., Krausmann, F., & Eisenmenger, N. (2010). Global patterns of materials use: A socioeconomic and geophysical analysis. *Ecological Economics*, 69, 1148-1158.
- Stern, D. (2004). The Rise and Fall of the Environmental Kuznets Curve. *World Development*, 32, 1419-1439.
- Stern, D.I., Common, M.S., & Barbier, E.B. (1996). Economic growth and environmental degradation: The environmental Kuznets curve and sustainable development. *World Development*, 24, 1151-1160.
- Stern, P.C. (1997). Environmentally Significant Consumption: Research Directions.
- Tallis, H.M., & Kareiva, P. (2006). Shaping global environmental decisions using socio-ecological models. *Trends in Ecology & Evolution*, 21, 562-568.
- Tebaldi, C., & Knutti, R. (2007). The use of the multi-model ensemble in probabilistic climate projections. *Philosophical Transactions of the Royal Society of London. Series A, Mathematical Physical and Engineering Sciences*, 365, 2053-2075.

- 669 Thomas, V., Theis, T., Lifset, R., Grasso, D., Kim, B., Koshland, C., & Pfahl, R.
 670 (2003). Industrial ecology: policy potential and research needs. *Environmental*
 671 *Engineering Science*, 20.
- 672 Tilman, D., Fargione, J., Wolff, B., D'Antonio, C., Dobson, A., Howarth, R.,
 673 Schindler, D., Schlesinger, W.H., Simberloff, D., & Swackhamer, D. (2001).
 674 Forecasting agriculturally driven global environmental change. *Science*, 292, 281-284.
- 675 UNEP. (2011). *Decoupling Natural Resource Use and Environmental Impacts from*
 676 *Economic Growth*.
 677 <http://www.unep.org/resourcepanel/Publications/Decoupling/tabid/56048>.
- 678 Utgikar, V.P., & Scott, J.P. (2006). Energy forecasting: Predictions, reality and
 679 analysis of causes of error. *Energy Policy*, 34, 3087-3092.
- 680 Valin, H., Sands, R.D., van der Mensbrugghe, D., Nelson, G.C., Ahammad, H., Blanc,
 681 E., Bodirsky, B., Fujimori, S., Hasegawa, T., Havlik, P., Heyhoe, E., Kyle, P., Mason-
 682 D'Croz, D., Paltsev, S., Rolinski, S., Tabeau, A., van Meijl, H., von Lampe, M., &
 683 Willenbockel, D. (2013). The future of food demand: understanding differences in
 684 global economic models. *Agricultural Economics*, 45.
- 685 van der Sluijs, J.P. (2002). A way out of the credibility crisis of models used in
 686 integrated environmental assessment. *Futures*, 34, 133-146.
- 687 Voet, E.V.D., Oers, L.V., & Nikolic, I. (2005). Dematerialization: Not just a matter of
 688 weight. *Journal of Industrial Ecology*, 8, 121-137.
- 689 Waggoner, P.E., & Ausubel, J.H. (2002). A framework for sustainability science: A
 690 renovated IPAT identity. *Proceedings of the National Academy of Sciences of the*
 691 *United States of America*, 99, 7860-7865.
- 692 Wiedmann, T. (2009). A review of recent multi-region input–output models used for
 693 consumption-based emission and resource accounting. *Ecological Economics*, 69,
 694 211-222.

- 695 Wigley, T.M.L., Clarke, L.E., Edmonds, J.A., Jacoby, H.D., Paltsev, S., Pitcher, H.,
 696 Reilly, J.M., Richels, R., Sarofim, M.C., & Smith, S.J. (2009). Uncertainties in
 697 climate stabilization. *Climatic Change*, 97, 85-121.
- 698 Wigley, T.M.L., & Raper, S.C.B. (2001). Interpretation of high projections for global-
 699 mean warming. *Science*, 293, 451-454.
- 700 Wise, M., Calvin, K., Thomson, A., Clarke, L., Bond-Lamberty, B., Sands, R., Smith,
 701 S.J., Janetos, A., & Edmonds, J. (2009). Implications of limiting CO₂ concentrations
 702 for land use and energy. *Science*, 324, 1183-1186.
- 703 Wright, S., & Muller-Landau, H. (2006). The future of tropical forest species.
- 704 *Biotropica*, 38, 287-301.
- 705 York, R., Rosa, E.A., & Dietz, T. (2003). STIRPAT, IPAT and ImPACT: analytic
 706 tools for unpacking the driving forces of environmental impacts. *Ecological*
 707 *Economics*, 46, 351-365.
- 708 Zhang, X., Davidson, E.A., Mauzerall, D.L., Searchinger, T.D., Dumas, P., & Shen,
 709 Y. (2015). Managing nitrogen for sustainable development. *Nature*, 528, 51-59.
- 710

711 **Table 1.** Summary of some key strengths and weaknesses of widely used large-scale approaches to forecasting global environmental change.

712

Method	Strengths	Weaknesses	Examples
Phenomenological models	Simple to parameterise and validate (at a high level); Suitable for top-down analysis of global or regional data; Easy to interpret.	Many embedded (opaque) assumptions; No explicit modelling of processes; Composite parameters are impossible to disaggregate.	Species Area Relationship; Environmental Kuznets Curve
Storyline scenarios	Intuitive to communicate; Maps readily to 'pathway' frameworks and socio-economic narratives; Captures 'snapshots' of continuous axes of discrimination (e.g., global vs regional, technological vs social).	Underestimate range of plausible future outcomes; Constrains thinking about alternative scenarios that cannot be accommodated across selected axes; Programmed with a fixed bundle of parameters.	Special Report on Emissions Scenarios; Millennium Ecosystem Assessment Report
Integrated Assessment Models	Based on well-verified economic methods for assimilating local to regional data; Aggregates results to produce 'bottom up' analysis of	Different storylines often borrow from same underlying models of drivers; Complex and heavily assumption driven; Difficult to	MiniCAM; MERGE; IGSM

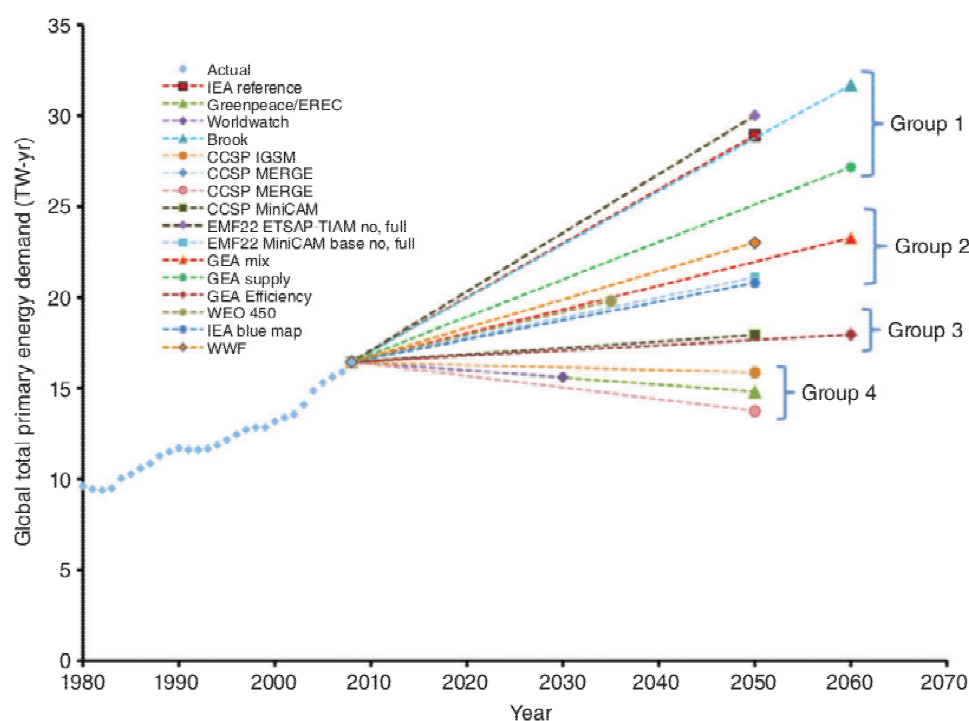
	change; Relatively blind to disciplinary borders; Can lead to probabilistic assessments.	determine sensitivities, especially in relationship to the constraints imposed by strong assumptions.	
Decomposition and Identity Approaches	Permits use of simple, bottom-up decompositions of aggregate drivers; Based on well-grounded methods developed in industrial ecology; Makes assumptions and exogenous inputs highly transparent; Contribution of each factor can be broken into fine-grained factors.	Rudimentary approaches have limited utility in forecasting; High-level aggregated parameters are often assumed rather than data-driven; Typically ignores problems of model selection/choice and stopping rules for 'sufficient' disaggregation are not clear.	ImPACT; STIRPAT
Global Climate (and Ecosystem) Models	Coupled (interlinked) system model of geo-physical and some biophysical processes; Captures interaction across multiple atmospheric and oceanic strata; Allow for forecasting using future forcing scenarios that are derived	Spatial grid-resolution makes simulation of fine-scale processes difficult; Simplified parameterization of poorly measured processes (e.g. clouds); Assumes hierarchical scaling of local-scale processes to biomes	HadCM3; CCSM; MAGICC; Madingley Model (GEM)

	from other modelling methods; Explicitly incorporates feedbacks.	and biosphere (GEM).	
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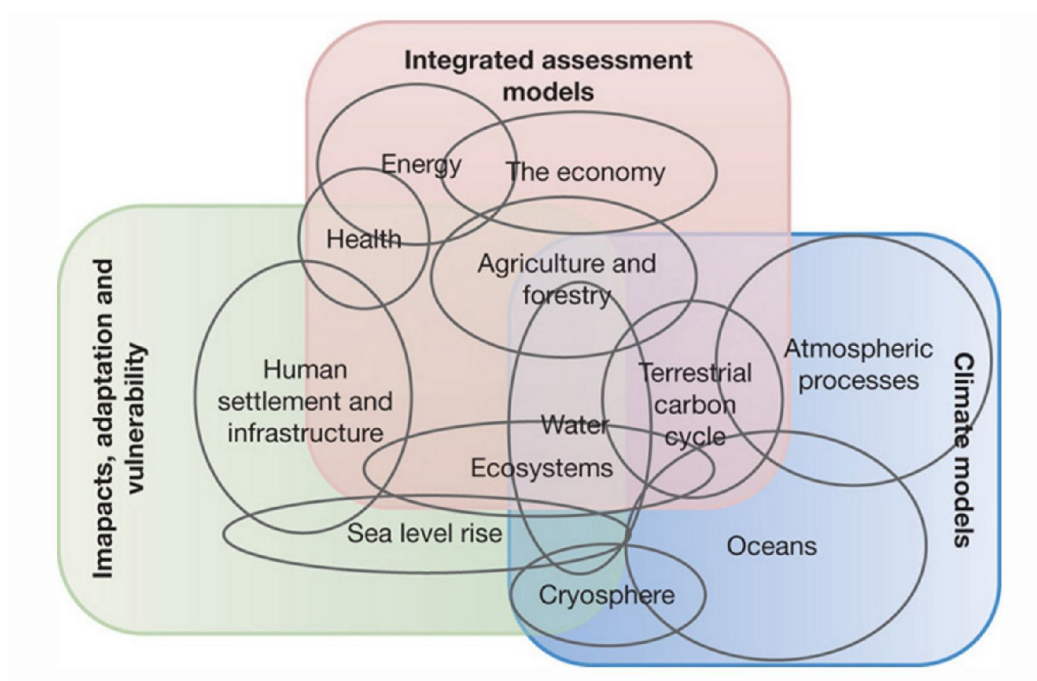
714 **Fig. 1.** Projected global energy demand trajectories for the 21st century, drawn from a
 715 wide range of storyline scenarios. Two notable points are that the results group into
 716 clusters (based on similar assumptions), but also that a wide range of possible futures
 717 can be imagined by groups working with different methodologies and goals. A major
 718 challenge of projecting change, beyond data and limitations, is coping with inherent
 719 uncertainties about future drivers of socio-economic decision-making.



720

721 *Source: Loftus et al. (2015)*

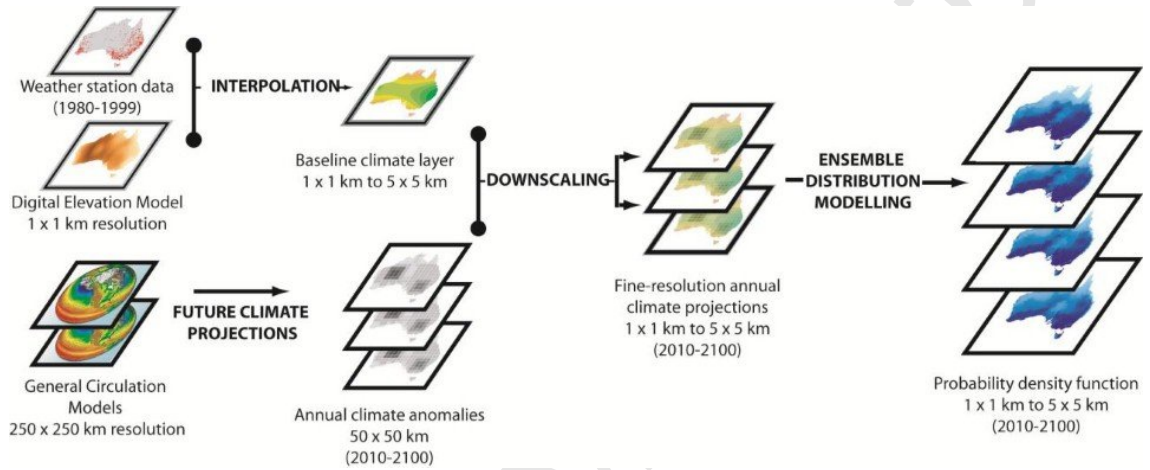
722
723 **Fig. 2.** Example of the multi-sectorial components of Integrated Assessment Models,
724 and how they link to assessments of environmental impacts and climate forecasts.



725
726 *Source: Moss et al. (2010)*

727
728

Fig. 3. Schematic depiction of ensemble forecasting of climate change, whereby high-resolution baseline climate grids from station data are linked to global climate models with good regional skill, to produce downscaled probabilistic multi-model predictions.



Source: Modified from Fordham *et al.* (2011)



