










# Use, calibration and verification of agroecological models for boreal environments: A review

Daniel Forster<sup>1</sup>  | Samuli Helama<sup>2</sup>  | Matthew T. Harrison<sup>3</sup>  |  
 Clarence Alan Rotz<sup>4</sup>  | Jinfeng Chang<sup>5</sup>  | Phillippe Ciais<sup>6</sup>  | Elizabeth Pattey<sup>7</sup>  |  
 Perttu Virkajärvi<sup>1</sup>  | Narasinha Shurpali<sup>1</sup> 

<sup>1</sup>Production Systems/Grasslands and Sustainable Farming, Natural Resources Institute Finland (LUKE), Maaninka, Finland

<sup>2</sup>Natural Resources Institute Finland (Luke), Rovaniemi, Finland

<sup>3</sup>Tasmanian Institute of Agriculture, University of Tasmania, Launceston, Tasmania, Australia

<sup>4</sup>USDA/Agricultural Research Service, University Park, Pennsylvania, USA

<sup>5</sup>College of Environmental and Resource Sciences, Zhejiang University, Hangzhou, China

<sup>6</sup>Laboratoire des Sciences du Climat et de l'Environnement, LSCE/IPSL, CEA-CNRS-UVSQ, Université Paris-Saclay, Gif-sur-Yvette, France

<sup>7</sup>Agriculture and Agri-Food Canada, Ottawa Research and Development Centre, Central Experimental Farm, Ottawa, Ontario, Canada

## Correspondence

Daniel Forster, Natural Resources Institute Finland (Luke), Halolantie 31 A, 71750 Maaninka, Finland.  
 Email: [daniel.forster@luke.fi](mailto:daniel.forster@luke.fi)

**Handling Editor:** Bernhard Schmid

## Funding information

Milk and Beef Production in Finland - Progressing Pathways Carbon-neutrality by 2035, Grant/Award Number: VN/28562/2020-MMM-2; Mechanisms for nitrous oxide uptake in cropping systems in different climate zones (NSINK), Grant/Award Number: 334422

## Abstract

Past assessments report negative impacts of the climate crisis in boreal areas; but milder and shorter winters and elevated atmospheric CO<sub>2</sub> may provide opportunities for agricultural productivity potentially playing a significant role in future food security. Arable cropping systems are expanding in boreal areas, but the regional mainstay will likely continue to be livestock production. Agroecological models can when appropriately calibrated and evaluated, facilitate improved productivity while minimising environmental impacts by identifying system interactions, and quantifying greenhouse gas emissions, soil carbon stocks and fertiliser use. While models designed for temperate and tropical zones abound, few are developed specifically for boreal zones, and there is uncertainty around the performance of existing models in boreal areas. We reviewed model performance across boreal environments and management systems. We identified a dearth of modelling studies in boreal regions, with the publication of three or less papers per year since the year 2000, constituting a significant research gap. Models IFSM and BASGRA\_N performed best in grassland production, DNDC best in predicting soil N<sub>2</sub>O and NH<sub>3</sub> emissions. No model outperformed all others, strengthening the case for ensemble modelling. Existing agroecological models would be worthy of further evaluation, providing model improvements designed for boreal systems.

## KEYWORDS

boreal region, carbon–nitrogen cycling, ecophysiological modelling, greenhouse gas emissions, soil carbon

## INTRODUCTION

Circumboreal countries adjacent to polar regions provide significant global greenhouse gas (GHG) emissions mitigation opportunities due to their potential for increased carbon sequestration and storage. Boreal soils are known to be important carbon stores and sources and may have the potential to act as further sinks (Lind et al., 2020). As the climate warms, boreal fertile soils,

previously either under permafrost or exposed to snow/thaw cycles, might release C either as CO<sub>2</sub> or methane and eventually become available for agriculture increasing the potential for global food production. Given the uncertainty around the risks and opportunities arising in boreal regions, agro-ecological simulation models may help elicit and contrast manifold metrics generated using management × environment scenarios (viz Christie et al., 2018, 2020). However, the vast majority of extant models

This is an open access article under the terms of the Creative Commons Attribution License, which permits use, distribution and reproduction in any medium, provided the original work is properly cited.

© 2022 The Authors. *Grassland Research* published by John Wiley & Sons Australia, Ltd. on behalf of Chinese Grassland Society and Lanzhou University.

have been designed for—and applied in—non-boreal areas characterised by milder temperatures falling within biological optima for plant growth and livestock performance (Harrison, Christie, et al., 2014; Harrison, Jackson, et al., 2014). As such, there is considerable uncertainty regarding the performance of these ‘temperate-borne’ models when applied in boreal conditions, for the accuracy in which agricultural productivity, C and N and GHG emissions are simulated.

The science underpinning ecosystems models has advanced over several decades into systems-based approaches for studying environmental processes as networks of ecophysiological interactions rather than in isolation (Jones et al., 2016). As such, models have increasingly become indispensable tools in both agricultural planning, adaptation and complementing measurements obtained from laboratory experiments or field trials. The use of models to simulate complex systems enables (1) integration and interpretation of multiple input variables, (2) prediction of the effects of climate change on whole landscapes and regions, (3) holistic quantification of GHG emissions and (4) improved mechanistic understanding of soil–plant–animal interactions, accounting for dynamic feedback loops and emergent properties (Harrison et al., 2012a, 2012b; Phelan et al., 2015). To some extent, models allow insight into interactions between processes and to make predictions and contrast alternative scenarios. Changes in soil carbon cycling due to management interventions for example can evolve over several decades (Forster et al., 2021); time horizons that are impractical and costly to measure in field experiments, and given the imminent need to reduce atmospheric GHGs (Alcock et al., 2015; Sándor et al., 2016, 2020), simulations permit assessment and exploration that would often be not otherwise possible (Harrison et al., 2017).

Agro-ecological models have long been used to simulate management × environment scenarios in Europe (Lugato et al., 2015; Metzger et al., 2005), North America (Guest et al., 2017; Khalil et al., 2020; D. Kim et al., 2019), Australia (Bilotto et al., 2021; Coleman et al., 1997) and elsewhere, with applications ranging from reductionist to holistic systems with scales ranging from the paddock to the planet (Harrison et al., 2011). In contrast, the number of model studies conducted in boreal climates—defined here as climates characterised by very cold winters and short, cool summers such as those found at latitudes greater than 60°—are much more limited. While process-based agro-physiological models have demonstrated reliability for the contexts in which they have developed, the majority of these models require (1) initialization, that is, input of initial data, (2) calibration or parameterisation, where simulations are calibrated by modifying model parameters/equations such that simulations better match measured variables and (3) evaluation, where calibrated models are compared with independently measured data and information (Harrison et al., 2019). There is limited evidence of either model calibration or evaluation for boreal environments, particularly in simulating GHG emissions. This key information gap forms a central focus of the present review.

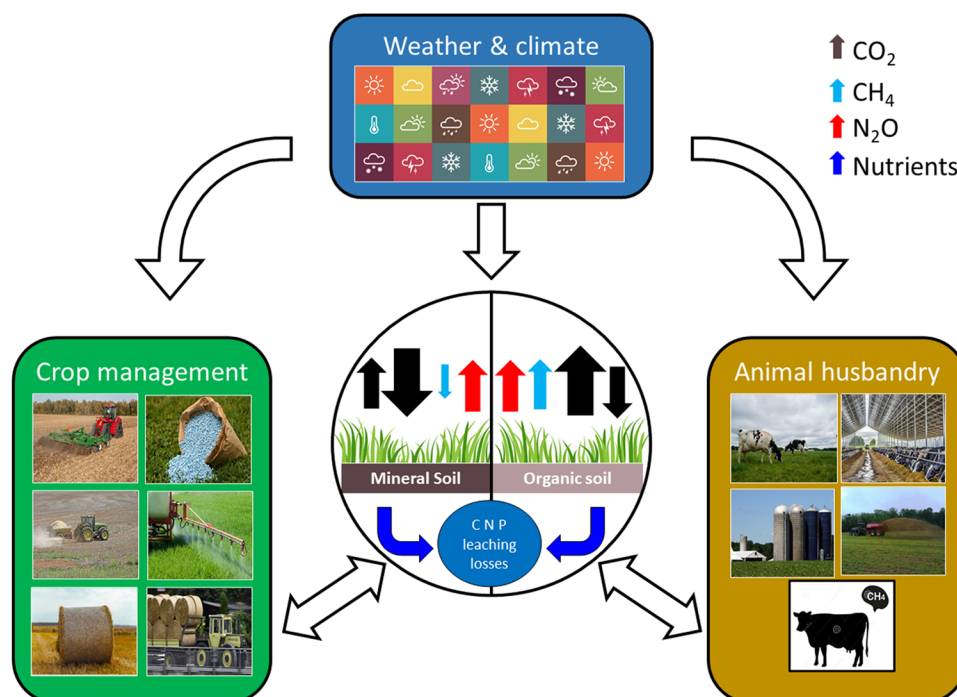
Studies have been carried out in diverse climatic zones, for example, Jose et al. (2016) examined the accuracy and efficiency of several models in simulating livestock-derived GHG's but paid less attention to specific environments, while Harrison, Christie, et al. (2014) and Harrison, Jackson, et al. (2014) and Ho et al. (2014) modelled the productivity, profitability and GHG emissions of farms as a function of herd management, animal genotype and pasture nutritive quality on farms in Australia. An ‘ensemble modelling’ approach was used by both Ehrhardt et al. (2018); and Sándor et al. (2020) to integrate multiple models and multiple management scenarios, although even the performance of the ensemble was limited in Scandinavian regions. Model intercomparisons have also been carried out for Canada (e.g., Guest et al., 2017; Sansoulet et al., 2014) with moderate success, and models have been successful in simulating agriculturally relevant outputs in a range of climates from Europe (Lugato et al., 2015; Metzger et al., 2005), North America (Höglind et al., 2016), Australia (Coleman et al., 1997), but studies over Boreal Europe are much more limited (Höglind et al., 2013; Korhonen et al., 2018).

Boreal regions will experience increasing environmental and agricultural pressure in the coming decades due to agricultural land use expansion and intensification (Unc et al., 2021). This pressure will mean that ecological and socioeconomic implications of farming become ever more important; models used in boreal conditions need to be able to deal with the contrasting soil and extreme seasonal climate conditions including snowfall (Jégo et al., 2014), the subzero temperatures as low as  $-35^{\circ}\text{C}$  to  $-45^{\circ}\text{C}$ , as well as site-specific livestock management in boreal ecosystems (Figure 1) if they are to be fit-for-purpose.

The purpose of this paper is to review the literature for process-oriented models dealing with managed grasslands in boreal countries, with an aim of identifying process deficiencies in model-based approaches for simulating agricultural and environmental variables, particularly productivity and GHG emissions. Identification of these deficits together with an assessment of the performance of existing models will then form the basis for recommendations for improving modelling of agro-ecosystems in boreal regions. Key areas examined include the performance of models in predicting biomass production, soil C cycling and GHG emissions from soils, plants, and livestock over the range of process-based models.

## GRASSLAND AGRICULTURE IN THE BOREAL REGION

The global livestock sector is responsible for about 8.1 Gt of  $\text{CO}_2\text{-e year}^{-1}$ ; about 16.5% of global anthropogenic GHG emissions (Twine, 2021) and 57% of that derived from food production (Xu et al., 2021). Livestock farming has been intensifying globally, including the conversion of natural grassland to pasture and increased  $\text{CH}_4$  and  $\text{N}_2\text{O}$  emissions, with warming effects on global temperatures being larger than the cooling effect of increased grassland carbon sinks (Chang et al., 2021).



**FIGURE 1** Interactions affecting soil/atmospheric fluxes of GHG's (coloured arrows) in agriculture in mineral and organic soils (central circle), including: climate inputs (blue box), crop management including sowing, fertilising, harvest, spraying, baling and transport (green box), animal husbandry including grazing, housing, feed, manure management and GHG emissions, both enteric and from subsequent manure (brown box). White arrows indicate interactions between compartments.

so that grasslands may become a net source of radiative forcing. While the growth in livestock numbers until 2050 will become increasingly underpinned by production of monogastrics (pigs and poultry), ruminant numbers through beef, sheep and dairy production systems will also continue to grow (Harrison et al., 2021).

The 26th Conference of the Parties (COP26) held in Glasgow (Glasgow Climate Pact, 2021) reaffirmed the terms agreed under Article 2 of the Paris Agreement (Paris Agreement to the United Nations Framework Convention on Climate Change, 2015), where more than 150 signatory countries pledged to make efforts to mitigate the impacts of climate change with the aim of keeping global temperatures well below 2°C above preindustrial levels. In the wake of this agreement there has been an increased interest in agricultural land use, and a recognition of the role of grasslands as potential carbon sinks (Whitehead, 2020), both in soils and in vegetation (Harrison et al., 2021). Carbon sequestration in the soil has been recognised as important for climate mitigation and enhancing food security and has led to the political target to increase soil carbon by 0.4%, or '4 per mille' per year (Minasny et al., 2017). According to the 6th Assessment Report (IPCC, 2021), the Paris Agreement targets are likely to be missed during the mid-21st century unless large reductions in GHG emissions are made in the coming decades. While there is hope that the development and application of carbon capture and storage technologies may develop in the future (Wilberforce et al., 2021), synthetic and industrial propositions for carbon capture are unlikely to provide the environmental and agricultural cobenefits offered by nature-based solutions (Harrison et al., 2021).

The need for sustained and harmonized efforts to mitigate global climate change become increasingly urgent. Opportunities arising from the poleward shift of agricultural activities, including increased cultivation and conversion from grassland to arable, and the implications for both food production and GHG emissions increases (Altdorff et al., 2021), have drawn relatively little academic attention, very likely missing opportunities for climate change mitigation and adaptation (IPCC, 2021). Circumboreal countries may have opportunities for increased sequestration efforts from increased crop-mediated CO<sub>2</sub> storage (Fan et al., 2019). Yet, they are also susceptible to climate change effects due to permafrost thawing (Hugelius et al., 2020). Boreal soils are known to be important carbon stores and may have the potential to act as further sinks from the enhancement of vegetation growth due to warming or as carbon sources from warming-induced decomposition of soil organic matter. In particular, peat soils—which occupy some  $3.7 \pm 0.5$  million km<sup>2</sup> (Hugelius et al., 2020) of boreal regions, storing  $415 \pm 150$  Gt C (Beaulne et al., 2021)—could become sources of carbon, including the potent CH<sub>4</sub>, due to warming-related increased soil respiration if management is not appropriately adapted to changing environmental conditions (Roulet, 2000). Northern peat soils being cultivated have also caused large carbon losses to the atmosphere in the history (Qiu et al., 2021) and in the present period (Carlson et al., 2017).

### Livestock grazing in boreal regions

In boreal areas, livestock farming on grasslands comprises a considerable part of the agricultural sector due to



limitations imposed by soil and climate, with permanent pasture comprising  $12\% \pm 8\%$  of Scandinavian,  $33\%$  of Canadian, and  $97\% \pm 2\%$  of agricultural land in Greenland and Iceland (FAOSTAT, 2022), thus livestock farming has an important role to play in mitigating agricultural GHG emissions in these regions.

Boreal economies are affected by climate change in both positive and negative ways. For example, elevated atmospheric  $\text{CO}_2$  concentrations stimulate photosynthesis of  $\text{C}_3$  vegetation and may result in longer growing seasons and increase production (Ergon et al., 2018), while the accompanying increase in microbial activity and shorter freezing periods due to climate warming will probably increase soil enzyme activity (Miura et al., 2020) and respiration (M. K. Kim & Henry, 2013), resulting in soil carbon losses, a problem of concern in soils of high organic content dominant in many boreal landscapes. Ecosystems resilience and vulnerability may also be influenced by changes to seasonal weather patterns, including increasingly variable regional precipitation, which can impact farm activities such as preserving hay and silage on which livestock rely during predictable dry periods (Chang-Fung-Martel et al., 2017; Harrison, Christie, et al., 2014; Harrison, Jackson, et al., 2014).

Since the 1970s, global grasslands have been estimated to transition from an anthropogenic sink to a source of anthropogenic GHG (Chang, Ciais, et al., 2015; Chang et al., 2021) due to increased livestock numbers and conversion of natural grasslands to improved pastures, according to a global modelling assessment. Nevertheless, boreal grasslands studies have concluded to either carbon sinks (Heimsch et al., 2021; Kätterer et al., 2011; Poeplau et al., 2015) or sources (Heikkinen et al., 2013; Lind et al., 2020; Lohila, 2004). Thus, a useful avenue of research is to reduce this uncertainty and assess if boreal grasslands are collectively either a net source or sink of  $\text{CO}_2$ , both historically and in the future. There remains considerable uncertainty due to regional variability even within boreal areas, so to adequately account for interregional differences, we discuss European and Canadian regions separately.

## The Nordic European region

Grasslands in the Nordic regions, which here include Iceland, Norway, Sweden and Finland (Figure 2) consist mainly of timothy (*Phleum pratense*) with red (*Trifolium pratense*) or white clover (*Trifolium repens*) (Helgadóttir et al., 2014). In Finland,  $10.4\%$  of the cultivated area (252 000 ha) is on organic soil (Myllis et al., 2012). This land may form significant sources of  $\text{N}_2\text{O}$  and  $\text{CO}_2$  when drained for agriculture, a problem that may persist even when cultivation ceases (Maljanen et al., 2010), making their careful management and preservation of critical importance. The managed grassland area in Finland has been steadily decreasing due to reductions in livestock farming (Niemi & Minna, 2019). Conversion to arable grassland is becoming common as agricultural technologies and practices advance and global warming makes temporary pasture and crop production more tenable (Soussana et al., 2010). Arable farmland in Finland is

generally thought to be a source of carbon emissions (Official Statistics of Finland, 2017). On the other hand, a study conducted in southern Finland using eddy covariance over 2 years showed that grasslands were a net sink of carbon (Heimsch et al., 2021). Contrarily, in Eastern Finland, mixed grassland of timothy and meadow fescue (*Festuca pratensis*) were found to be a large carbon source when offtakes were taken into account (Lind et al., 2020), although to date GHG emissions associated with grasslands in Finland have not been well quantified, and large-scale carbon balance estimates are not available.

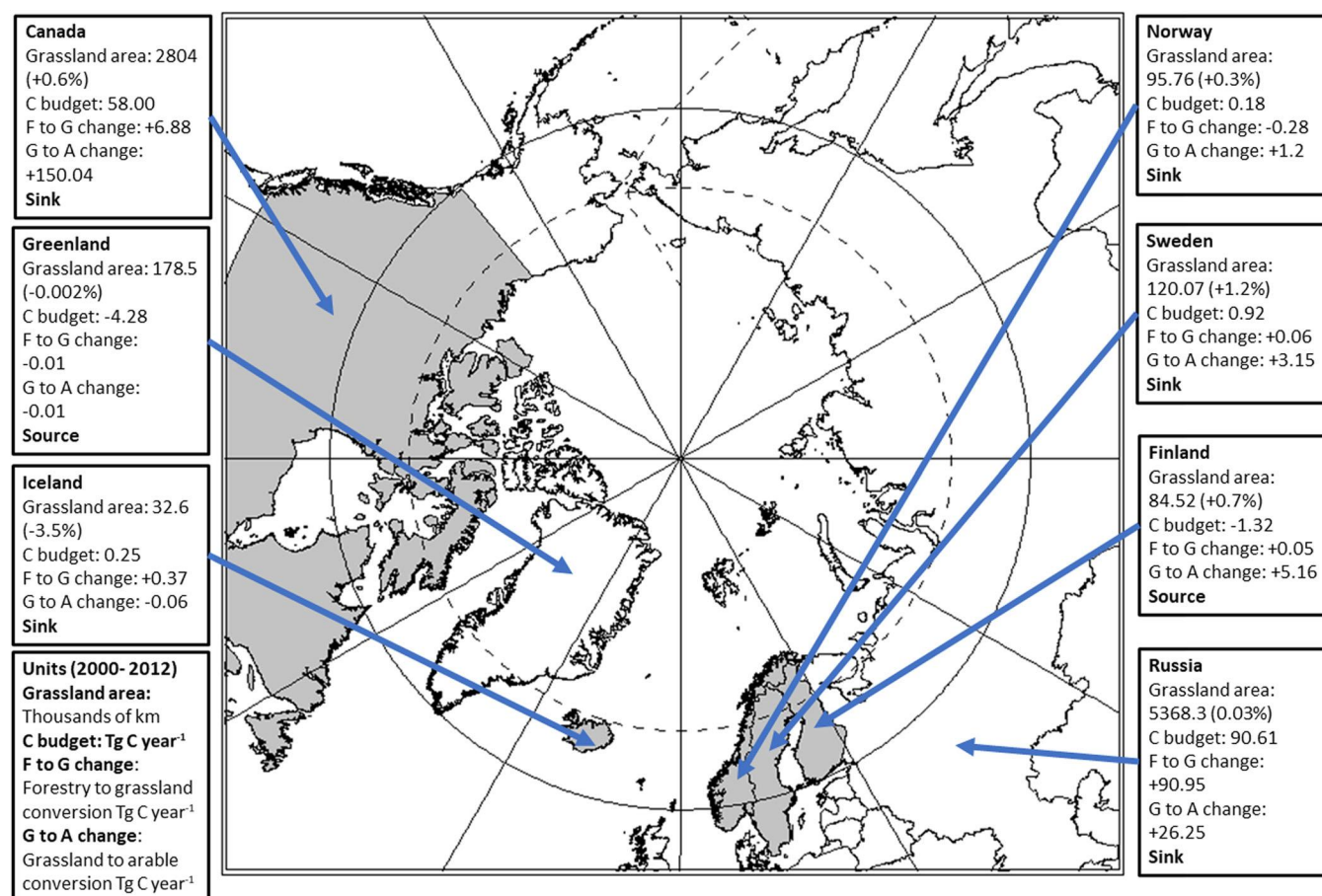
In Sweden, agricultural land comprises around 3 232 039 ha,  $7.9\%$  of the surface area (Lundblad, 2015), of which  $16\%$  is pasture, and  $7\%$  (225 722 ha) is on organic soil. There is good potential to increase soil organic carbon (SOC) stocks further with the addition of ryegrass cover cropping methods (Poeplau et al., 2015). In Norway, while the bulk of carbon storage is thought to be in forests and wetland, agricultural areas have been envisioned as being carbon sources (de Wit et al., 2015). Bartlett et al. (2020) estimated grassland SOC storage in Norway to be  $98 \text{ t C ha}^{-1}$ , although there is a lack of data on deeper soil C accumulation and in grasslands more generally. Grasslands in Iceland have been found to act as C sinks due to elevated N inputs both from agriculture and sea-bird manuring waiving N as a limiting factor (Leblans et al., 2017).

## Canada

In Canada, timothy (*Phleum pratense*), alfalfa (*Medicago sativa*), Kentucky bluegrass (*Poa pratense*), crested wheatgrass (*Agropyron cristatum*) and tall fescue (*Festuca arundinacea*), all of which have good winter hardiness and are suited to cooler climates, are usually grown for feeding livestock. A study of the sink potential of agricultural soils (Boehm et al., 2004) indicated enhanced C sequestration associated with conversion to zero tillage, and a study in Canada using Roth-C (Fan et al., 2019) indicated that increased plant inputs made cultivated soils a sink of carbon on the whole. However, this sink is steadily decreasing since then. Moreover, forage areas are increasingly converted to cropland (Mardian et al., 2021) and may become a net source of carbon if such trends continue due to (1) soil cultivation enhancing soil carbon turnover, (2) shallower root profiles and reduced soil carbon buried deeper in the profile of annual crops relative to perennial, deep-rooted pastures and (3) increased use of nitrogen fertilizers, which may increase emissions of  $\text{N}_2\text{O}$  (Harrison et al., 2016; Shcherbak et al., 2014) while enhancing carbon sequestration.

## METHODS

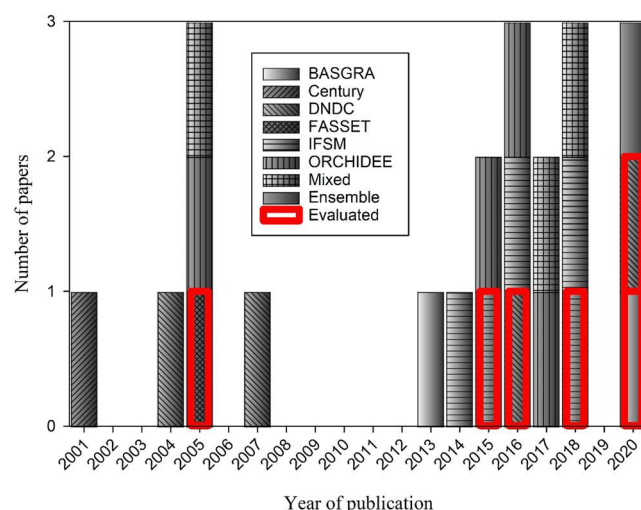
Independent literature searches were conducted using the Web of Science Database® and Google Scholar between May 4 to July 30, 2021. Search terms were: (a) grasslands, modelling and GHG emissions, soil carbon, methane,



**FIGURE 2** Regions examined in this study (greyed areas) and other boreal regions not included in this study (in white). Total grassland area (thousands of km) and percent change, carbon budget, and change associated with conversion from forestry to grassland (F to G change) and grassland to cropland (G to A change) in Tg C year<sup>-1</sup>, for the period 2000–2012 modelled using ORCHIDEE\_GM (Chang et al., 2021), and a summary carbon ‘Sink’ or ‘Source’, in bold.

carbon dioxide, nitrous oxide accompanied by words restricting results to boreal foci: boreal, Finland, Sweden, Norway, Denmark, Iceland, Canada and Russia. Our searches for Icelandic, Estonian and Latvian grasslands studies did not result in any literature reference identified. Only original research articles (not reviews or meta-analyses) were included. Papers were excluded if they had primary focus on statistical modelling, measurements or monitoring without modelling, if they did not represent regions within scope of the present review, for example, Scandinavia, Iceland, Canada, or those not published in English. No papers from boreal Russia, in theory a major area for this type of research, were found. Excluding duplicate hits, the searches provided an initial total of 65 papers published between 1998 and 2020, of which on inspection 21 were retained as meeting our criteria for inclusion. Publications were weighted towards recent studies as 10 were published in or after 2016, three of which were published in 2020, and no papers between 2008 and 2012 were found (Figure 3).

For an assessment of model validation results, an assessment method was adapted from Despotovic et al. (2016), He et al. (2020), and Moriasi et al. (2015): model performance was graded as ‘poor’, ‘fair’, ‘good’ or ‘excellent’. Validation results from each article were extracted and scored from 1 to 4 on this scale (see Supporting Information material for details), and mean



**FIGURE 3** Number and type of model used in boreal modelling studies by year. Ensemble and mixed model approaches incorporate various models, but these approaches are judged sufficiently different from single model methods to be classed as models in their own right. Red highlighted papers contain model evaluations.

scores across methods were calculated. A final score was calculated as the sum of the validation scores. It is recognised that this methodology has its weaknesses, in that it attempts to compare the relative merits of a range

of validation equations over a broad suite of variables, regions and timescales, using subjective interpretations of the results found, though every effort has been made to ensure our results reflect those of the original authors.

## RESULTS

Overall, seven models were found, though one, CATIMO, only appeared in ensemble studies and so is treated as such. The key result in Figure 3 is the dearth of modelling studies in Boreal grasslands with three or less studies per year in the last 20 years, in contrast with a large number of publications on boreal forests. This result is investigated further in the following sections as each model in the literature search is discussed.

### BASGRA and BASGRA\_N

The BASGRA\_N model is an advancement on the earlier BASGRA grasslands model (Höglind et al., 2013, 2016, 2020); both models are designed for cold climate growth of timothy in Scandinavia. The model is process-based and uses the source/sink concept for net exports and imports of photosynthetic assimilates between photosynthetically developed leaf tissue and source tissue, developing plant parts and roots (Persson et al., 2019) on a daily time-step with temperature and day length as driving variables. The model is also designed to account

for winter conditions such as snow cover and soil freezing which are common to boreal regions and typically lead to high rates of tiller death and reduced spring season growth. BASGRA is the only model which simulates tiller death and growth dynamics in timothy, making it especially useful in northern regions where this is the primary forage grass. To date, there is a lack of comparable Scandinavian data from which to assess BASGRA\_N model performance regarding tiller dynamics in response to N fertilisation (the fraction of vegetative, nonelongating and elongating generative tillers in swards and their winter survival) as well as fluxes in C and N (soil respiration, N<sub>2</sub>O-emissions and N leaching), making this an area in need of further research.

BASGRA was designed with simulating biomass production and feed value, and accordingly included all the processes of tillering, foliar dynamics, winter death and spring regrowth of timothy. The aim of the updated BASGRA\_N model (Höglind et al., 2020) was to supplement this with a number of additional features such as forage nutritive values, C and N cycling and GHG emissions, as well as accounting for N limiting conditions and the dynamics of cell wall content. Otherwise, the updated model retains many of the key features of its predecessor (Höglind et al., 2013). Model performance assessment was only assessed by Höglind et al. (2020) who reported reasonable performance in simulating dry matter production and good performance for crude proteins, but less reliable correlations between fertilisation and dry matter production (Table 1).

TABLE 1 Summary of score given for each model for variables validated

Variable	BASGRA_N	CATIMO	DNDC	DNDC v.CAN	FASSET	IFSM	STICS
<b>Biomass</b>							
CP (g m <sup>-2</sup> )	Good	-	-	-	-	-	-
DM (g m <sup>-2</sup> )	Good	Fair	Fair	-	-	Good	Fair
LAI	Good	Excellent	-	-	-	-	Excellent
NDF	-	-	-	-	-	Good	-
N (uptake)	-	-	-	-	-	Fair	-
P (uptake)	-	-	-	-	-	Good	-
N fertilizer DM	Poor	Poor	-	-	-	-	Fair
<b>Gas emissions</b>							
N <sub>2</sub> O	-	-	Fair	-	Good	-	-
NH <sub>3</sub>	-	-	Fair	Good	-	-	-
<b>Soil climate</b>							
WFPS	-	-	Fair	-	-	-	-
Soil temperature (°C)	-	-	Good	-	-	-	-
N (mineral) (kg ha <sup>-1</sup> )	-	-	Poor	Good	-	-	-
Number of studies	2	1	2	1	1	1	1
Overall model score	10	7	12	6	3	11	8

Note: Values represent the mean of scores given to each validation method where more than one was used to assess the variable. Validation data was extracted from five papers (Chatskikh et al., 2005; Congreves et al., 2016; He et al., 2020; Höglind et al., 2020; Jégo et al., 2015; Korhonen et al., 2018) from the initial literature search and summarised according to a system derived from Despotovic et al. (2016); He et al. (2020); and Moriasi et al. (2015).



## CENTURY and DayCent

The CENTURY model is used to assess SOM processes involving C, N, P and S, and the effects of changes in management to soil processes. The model consists of a variety of submodels which can simulate grasslands, agricultural crop systems, forest systems and savannahs and works on a monthly time-step and is useful for modelling long-term changes to SOM of over 10–100 years. DayCent is a daily timestep version of CENTURY (Del Grosso et al., 2001) developed to link atmospheric models to better estimate trace gas fluxes from different environments (Parton et al., 1998). For boreal regions, Smith et al. (2001) simulated the effects of management changes resulting in increases of sequestered C of 0.62 and 0.44 Mg C ha year<sup>-1</sup>, respectively for conversion of arable to grassland and the inclusion of forage in crop rotations in western Canada using DayCent.

## DeNitrification-DeComposition (DNDC) and variants

The field-based DNDC model was designed as a process-based model which computes a range of descriptors related to soil carbon storage such as estimated soil-atmosphere fluxes, soil organic matter (SOM) storage and C and N related biogeochemistry in agricultural systems by tracking groups of microbes active under different environmental conditions as well as predicting crop and livestock yields (EOS, 2017). DNDC has been used with some success in a range of boreal grassland situations, although the main model was primarily used in arable systems. DNDC has produced a suite of models adapted to different region and management practices identified by a prefix. For example, NZ-DNDC is a New Zealand-specific model, and the Manure DNDC model specialises in manure emissions from livestock farms. Unless specified with a prefix we refer to the main crop model in this review.

Grant et al. (2004) used DNDC to estimate the impact of change in management practices on N<sub>2</sub>O emissions for seven major soil regions in Canada from 1970 to 2029. They estimated significant reduction of N<sub>2</sub>O emissions by converting from arable to managed permanent grassland and adopting no-tillage (Table 2). However, other factors such as soil type and regional differences influenced the general trend, like the adoption of no-tillage in eastern Canada, which would increase N<sub>2</sub>O emissions as soils are wetter than in the semi-arid zone of the Prairies. However, the conversion of cropland to grassland in eastern Canada would strongly reduce N<sub>2</sub>O emissions because of depleted organic matter of the mineral soils and reduced N fertilisation. This overall decrease was attributed to the slower decomposition of SOM in no-till compared to conventional tillage. Increasing N-fertilizer application rates by 50% would increase average emissions by 32%, while decreasing them by 50% would decrease emissions by 16%.

He et al. (2020) calibrated and validated DNDC to simulate the impacts of manure management practices,

including slurry application rates and seasonal timing, on N<sub>2</sub>O emissions, using N<sub>2</sub>O flux data, soil moisture, soil inorganic N, biomass and soil temperature measured from managed grasslands in contrasting climates of Canada (Table 2). The performance of DNDC was evaluated using NRMSE and NARE (See Supporting Information material), statistics, and was found ‘fair’ (<20%) to ‘excellent’ (≤10%) for predicting biomass, ‘good’ (<10% to ≤20%) for predicting soil temperature, but only ‘acceptable’ (<20%) in estimating soil water and inorganic N contents (see Table S2). The latter result could be associated with the limitations of a cascade water transfer sub-model and inaccuracies in simulating root development/uptake. The DNDC model demonstrated ‘fair’ (NRMSE <20% to <30%) performance in predicting daily N<sub>2</sub>O fluxes, capturing the impact of the timing and rate of slurry application and soil texture on total N<sub>2</sub>O emissions. Finally, the authors recommended improvements to DNDC simulation of soil freeze-thaw cycles, manure decomposition dynamics, soil water storage, rainfall canopy interception, and microbial denitrification and nitrification activities in grasslands. A modified version called DNDC v.CAN was used by Congreves et al. (2016) in Quebec, Canada, to simulate the effects of swine slurry distribution on grasslands and bare soil, improving the models capacity to predict NH<sub>3</sub> losses from soils, and also provided reasonable model validation results (Table S2). Levy et al. (2007) conducted a study for all of Europe in which grasslands in a range of biogeographical regions were assessed, including arctic and boreal regions. Their study did not attempt to model changes in atmospheric CO<sub>2</sub> or temperature resulting from climate change and concluded that mitigating livestock-based GHG emissions rather than attempts to increase C sinks was more likely to be beneficial in terms of limiting global temperature change.

## FASSET

FASSET is a whole farm soil–plant–atmosphere model developed by the Danish Institute of Agricultural and Fisheries Economics to deal with all major N flows at the farm level (Jacobsen et al., 1998). FASSET dynamically simulates the soil-plant-climate system at the field level on a daily time-step (Olesen et al., 2002). This was further developed by Chatskikh et al. (2005), who developed and added an algorithm for N<sub>2</sub>O production and emission from agricultural soils. The model of Chatskikh et al. (2005) simulated carbon and nitrogen turnover on a daily basis. Both nitrification and denitrification were included in their model as sources for N<sub>2</sub>O production with N<sub>2</sub>O emissions influenced by soil microbial and physical conditions. Their model was tested on experimental data of N<sub>2</sub>O emissions from grasslands in Finland, Denmark and UK, thus at sites that differed in their climatic conditions, soil properties and management. The simulated N<sub>2</sub>O emissions showed a nonlinear response to increasing N rates with increasing emission factors at higher N rates. This result aligns with trends reported in

TABLE 2 Summary of papers examined in this review

Serial	Model	Reference	Location	Simulated period	Management	Intervention	Outcome
1	BASGRA	Höglind et al., 2013	Northern Europe, Russia	1960–1990 and 2040–2065	Permanent pasture	Irrigated and non-irrigated pasture	Increased DM yields: 14% irrigated and 11% non-irrigated.
2	BASGRA_N	Höglind et al., 2020	Norway	1985–1988	Permanent pasture	Model validation for dry matter, crude protein, NDF and soil N under different fertiliser regimes	Yield and CP responses to N modelled satisfactorily.
3	CENTURY	Smith et al., 2001	Canada	2000–2010	Arable and pasture	Conversion of arable to pasture, crop rotations and variable fertiliser applications	Conversion of arable to pasture, conversion to forage and conversion to no-till and reduction of summer fallow period may sequester 0.62, 0.44, 0.13 and 0.15 Mg C ha <sup>-1</sup> year <sup>-1</sup> respectively.
4	DNDC	Grant et al., 2004	Canada	1970–2029	Land-use conversion	Conversion of arable to pasture and conversion to no-tillage	60% and 33% reduced N <sub>2</sub> O emissions from conversion to grass, and conversion to no-tillage respectively.
5	DNDC	He et al., 2020	Canada	2001–2003	Grassland	Manure application rate (zero, spring, autumn, split between spring and fall) and timing (early spring, late spring, split)	Daily soil N 2.2–45.8 kg N ha <sup>-1</sup> , cumulative soil N 44.4–1327.3 kg N ha <sup>-1</sup> .
6	DNDC	Levy et al., 2007	Europe-wide	20 years	Grasslands	Business as usual over 20 years	Most grasslands estimated as net sources of GHG's, N <sub>2</sub> O and CH <sub>4</sub> (23 Tg C year <sup>-1</sup> ). Suggested enteric CH <sub>4</sub> reduction more effective overall than sequestration attempts.
7	DNDC v.CAN	Congreves et al., 2016	Quebec, Canada	1999–2005	Grassland and bare soil.	Swine slurry applied to grassland and bare soil	Reduced errors for DNDC v.CAN compared to DNDC for mineral N loss.
8	Ensemble	Sándor et al., 2020	Canada, France, India, UK	2002–2012	Permanent pasture	Grazed long-term grasslands	GPP: 1763 g C m <sup>-2</sup> year <sup>-1</sup> , RECO: 1561 g C m <sup>-2</sup> year <sup>-1</sup> , NEE: 610 to 66 g C m <sup>-2</sup> year <sup>-1</sup>
9	FASSET	Chatskikh et al., 2005	Denmark, Finland, UK	1981, 2000–2002 and Feb–Nov 2002	Grassland	Lawn with control, N fertiliser and slurry treatments, annual N at 200–250 kg N ha <sup>-1</sup> , and conventional vs organic pasture with different fertiliser applications	N <sub>2</sub> O emissions model was able to predict seasonal patterns and temporal variability on N <sub>2</sub> O patterns in the three sites. Non-linear relationships between N inputs and N <sub>2</sub> O emissions and sensitivity to fertiliser type and soil texture found.
10	IFSM	Alemu et al., 2016	Canada	2004–2006	Dairy pasture	Control, single application (49.2 ± 8.0) and split application [2 × (48.7 ± 4.3)] of hog manure	Annual CH <sub>4</sub> emissions for control, single and split applications were 18.1, 25.9 and 26.5 kg CO <sub>2</sub> -e per kg liveweight of stock.
11	IFSM	Cordeiro et al., 2019	Canada	1990–2016 and 2020–2079	Dairy pasture	Three time periods: reference (1990–2016), near future (2020–2049) and distant future (2050–2079)	Expansions of cropland resulted in increased grass-legume production of 8% to 52% and total production increase of 11% to 105%.
12	IFSM	Duchemin et al., 2019	Quebec, Canada	1986–2015	Pasture, switchgrass	NA	Average DM yields were between 9.6 and 11 t ha <sup>-1</sup> .
13	IFSM	Jégo et al., 2015	Canada	1991–2011	Pasture	Dry matter yields across contrasting climate zones	DM simulated (2.94 ± 0.74) vs. simulated (2.77 ± 0.52) t ha <sup>-1</sup> , and NDF simulated (0.42 ± 0.04) vs simulated (0.43 ± 0.02) g g <sup>-1</sup> DM.

(Continues)



TABLE 2 (Continued)

Serial	Model	Reference	Location	Simulated period	Management	Intervention	Outcome
14	IFSM	Thivierge et al., 2017	Canada	1971–2000 and 2020–2049	Dairy production systems	Impact of climate change on range of agro-economic parameters	Increased NH <sub>3</sub> (+18% to +54%), increased manure stored methane (+26% to +120%). Variable projected N footprint (–15% to +46%) and C footprint (–5% to +9%).
15	ORCHIDEE	Ciais et al., 2005	Europe-wide	2003	Various grassland, forestry, arable	Drought and ordinary year comparison	GPP reduced 30% and increased CO <sub>2</sub> emissions (0.5 Pg C year <sup>–1</sup> ) under drought conditions.
16	ORCHIDEE_GM	Chang, Ciais, et al., 2015	Europe-wide	1961–2010	Fertilised grassland soils	Intensive and extensive grasslands	Net C balance at the continental scale was estimated to be a net sink of 15 ± 7 g m <sup>–2</sup> year <sup>–1</sup> .
17	ORCHIDEE_GM	Chang et al., 2016	Europe-wide	1991–2010	Grasslands	Intensive and extensive grasslands	Net biome production increased by 24%–31% resulting from increased atmospheric CO <sub>2</sub> and reduced management intensity.
18	ORCHIDEE_GM	Chang et al., 2017	Europe-wide	1901–2100	Grasslands	Intensive and extensive grasslands	Net NPP increase projected and attributed to rising CO <sub>2</sub> .
19	Mixed model	Korhonen et al., 2018	Northern Europe, Canada	1991–2007	Pasture, Timothy	Two & three N application rates, early vs late cutting	Models tended to underestimate yield of first cut (BASGRA_N, STICS, CATIMO).
20	Mixed model	Desjardins et al., 2005	Canada	1980–2030	Pasture, arable	Converting cropland to rangeland	Potential 6% increase in soil C after conversion to rangeland. 200 Mg CO <sub>2</sub> e ha <sup>–1</sup> reduction.
21	Mixed model	Guest et al., 2017	Canada	1993–2007	Spring wheat	Model comparison	10% underestimation of soil N for DayCent, 22% for STICS, 1% for DNDC.

a global meta-analysis by Shcherbak et al. (2014). Increased soil clay content increased the simulated emissions, which were further increased at higher temperatures but generally decreased by increasing annual rainfall. Slightly higher emissions from grazed grasslands, compared with cut grasslands at similar rates of total N input (fertiliser and animal excreta) were evident. Validation results for N<sub>2</sub>O emission simulations were good for Pearson's correlation and reasonable, though model efficiency was less satisfactory (Table S3). The authors concluded that there was greater potential for reducing N<sub>2</sub>O emissions for intensively grazed grasslands on fine-textured soils, compared with the N<sub>2</sub>O mitigation potential of extensively mown grasslands on sandy soils.

## IFSM

The Integrated Farm System Model (IFSM) was designed to simulate all components of dairy or beef production at the whole farm level (Rotz et al., 2018) and uses discrete sub-models to simulate inputs and forecast costs and outputs. The model can simulate production of grass, small grain, soybean, alfalfa, and corn cropping systems, including growth, harvest and storage, as well as dairy outputs. It also has a beef component to simulate beef production systems. IFSM simulates environmental metrics including soil and animal-derived GHG emissions, NH<sub>3</sub> volatilisation and NO<sub>3</sub> and P leaching and

runoff, and so has the potential to identify management approaches to mitigate GHG emissions and determine reductions in the C footprint of farm products. Most studies using IFSM have been carried out in the United States and Canada, with the bulk of these being carried out in temperate or maritime climate conditions. This suggests that some agricultural GHG emissions models may not be well used outside their region of development, similar to the finding by Ara et al. (2021) regarding development and use of agricultural decision-support systems.

A study on switchgrass (*Panicum virgatum*) in southern Quebec was carried out by Duchemin et al. (2019) where biomass production was predicted reasonably well despite not being designed to work with this type of grass. A larger study in Canada (Thivierge et al., 2017) used the IFSM model in conjunction with climate models (CanESM2, CanRCM4 and HadGEM2) to project the impacts of climate change on agronomic and environmental performance. Projected climate effects resulted in increased production of grasses but also increased NH<sub>3</sub> and CH<sub>4</sub> emissions from manure storage, although there were limited reductions of CH<sub>4</sub> emissions from enteric and field applied manure.

IFSM has also been used in Newfoundland (Canada) to model climate change impacts on dairy farm production until 2079 (Cordeiro et al., 2019). Projected climate tended to increase production, particularly in grass and legumes (Table 2), but reduced farming operation opportunities due to increased rainfall during the

growing season. This study focussed on production and did not examine GHG emissions or other environmental effects. A study by Jégo et al. (2015) explored this aspect using data on timothy and alfalfa grown in Canada, for sites in New Brunswick, Quebec and Alberta. The IFSM successfully represented forage production and neutral detergent fibre (NDF) concentrations under these conditions, although N uptake required larger sample sizes to reduce uncertainty.

Alemu et al. (2016) evaluated the impact of time and amount of hog manure application on farm productivity and GHG emissions from a Canadian cow-calf production system using the CCM (Coupled Components Model) and IFSM. They found that around 75% of the total farm GHG emissions were from enteric CH<sub>4</sub> production, similar to other livestock production systems (Christie et al., 2014; Harrison et al., 2016; Rawnsley et al., 2018; Taylor et al., 2016). Moreover, application of hog manure on grassland showed a mean GHG emission increase, mainly from enteric CH<sub>4</sub> and soil N<sub>2</sub>O emissions for waste management scenarios. The authors concluded with the need of further component and whole-farm assessments to better understand the impact of the amount and timing of livestock manure application on GHG emissions from beef production systems. Validation results for dry matter and NDF has shown adequate model performance, although N uptake simulation was only 'fair' (Table S4). This result aligns with other model intercomparison studies on the ability of systems models to simulate N cycling under field conditions (Bilotto et al., 2021).

## ORCHIDEE\_GM

In a study spanning the entire Western European region, Chang, Viovy, et al. (2015) estimated the GHG balance (CO<sub>2</sub>, CH<sub>4</sub> and N<sub>2</sub>O). They used the process-based biogeochemical model, ORCHIDEE-GM over the period 1961–2010 (Table 2). This model (Chang, Ciais, et al., 2015; Chang et al., 2013) is an evolution of the ORCHIDEE land surface model (Ciais et al., 2005; Krinner et al., 2005; Piao et al., 2007) that includes a better representation of grassland phenology as well as of management, from the PaSim grassland model (Riedo et al., 2002). The management module includes grazing during a suitable season, and cutting for forage production, assuming that production of grass production is high enough, farmers will use a fraction of each grid cell for grazing and the other for forage production to feed animals in winter. The model can be forced by spatially explicit animal density and weight maps or calculate the maximum capacity to use all the production. CH<sub>4</sub> and N<sub>2</sub>O emissions are represented by emission factors, not by mechanistic equations.

At the farm scale in Europe, the net C balance was roughly halved down to a small sink or nearly neutral flux of 8 g C m<sup>-2</sup> year<sup>-1</sup> in ORCHIDEE\_GM simulations from 1961 to 2010. At the continental scale, adding CH<sub>4</sub> and N<sub>2</sub>O emissions to determine net

ecosystem exchange, grasslands remained a net GHG sink as the CO<sub>2</sub> sink offset N<sub>2</sub>O and grazing animal CH<sub>4</sub> emissions, however, at the farm scale these additions resulted in a net source of -50 g C-CO<sub>2</sub>e m<sup>-2</sup> year<sup>-1</sup>. The largest net GHG sink by grassland was found in the British Isles. However, here too the GHG balance shifted to a net source when considered at the farm scale. The authors concluded that enhanced GHG balance reflects the combination of a positive trend of net primary production due to CO<sub>2</sub>, climate and nitrogen fertilization and the diminishing requirement for grass forage due to the Europe-wide reduction in livestock numbers. Another study by Chang, Viovy, et al. (2015) estimated changes in potential productivity and potential grass-fed ruminant livestock density across European grasslands including boreal areas over the period 1961–2010. When compared with agricultural statistics (Eurostat and FAOstat), ORCHIDEE-GM gave a good reproduction of the regional gradients of annual grassland productivity and ruminant livestock density, although the model tended to systematically overestimate the absolute values of productivity in most regions, suggesting that most grid cells remained below their potential grassland productivity due to possible nutrient and biotic limitations on plant growth. Another work showed that net biome production (NBP) in the Nordic region increased between 1991 and 2010, mainly in response to changes in climate and atmospheric CO<sub>2</sub> concentration but also partly as a result of reduced management intensity and livestock numbers (Chang et al., 2016). And another study projected increased net primary productivity (NPP) and earlier and longer grazing periods under higher warming levels (Chang et al., 2017).

## Model intercomparisons and ensemble modelling

Desjardins et al. (2005) investigated the role of agriculture in influencing the GHG budget and examined the opportunities from management practices for increasing soil C sequestration in Canada (see Table 2). Simulations of C and N cycling by CENTURY and DNDC models were carried out for five locations across Canada over a 30-year period to outline the potential trade-offs between C sequestration and increased N<sub>2</sub>O emissions. Simulations suggested that conversion of cropland to grassland would result in the largest reduction in net GHG emissions. Moreover, nutrient additions, that is, by fertilizers would result in a small increase in GHG emissions. Finally, improved growing conditions would increase the soil C by 6% over the more recent 15-year period, as demonstrated for one of their sites in Alberta. The DNDC model was compared with DayCent, and the soil-atmosphere-crop model STICS (Brisson et al., 2003) in field studies in eastern Canada (Guest et al., 2017), and was the most reliable of the three at predicting soil N (-1%), although all three performed better predicting soil moisture and evapotranspiration than other output variables. STICS exhibited the largest underestimation (22%) for predicting soil N. The STICS model was recently adapted to eastern Canada conditions, by

calibrating relevant cultivars and adding a snow module (Jégo et al., 2014) as it was mostly developed and used in temperate and more tropical regions than in northern ones.

Korhonen et al. (2018) used BASGRA to study the production of timothy across Finland, Sweden, Norway and Canada alongside two other models, STICS and the timothy biomass production model CATIMO (Canadian Timothy Model) (Bonesmo & Bélanger, 2002), although these two did not appear anywhere else in the literature search. All three models performed similarly, none produced accurate simulations of grass growth. STICS simulated aboveground spring growth better than the other two models, though BASGRA was better for summer growth (Tables S1, S5 and S6), and was the most accurate overall, but would benefit from additional soil descriptive data input.

Combined with previous estimates of CO<sub>2</sub> emissions by Smith et al. (2001) who used the CENTURY model, Grant et al. (2004) postulated that the most beneficial management practices for reducing both N<sub>2</sub>O and CO<sub>2</sub> emissions is the conversion from conventional tillage of croplands to permanent grassland, the implementation of reduced tillage, and the reduction of summer fallow. A trade-off in GHG flux with greater N<sub>2</sub>O emissions and a comparable increase in C storage would be reached when 50% more N-fertilizer was added.

Ensemble modelling approaches have also been used in recent years to reduce the generalised error of predictions, for example, Ehrhardt et al. (2018); Farina et al. (2021); Sándor et al. (2020, 2018). Sándor et al. (2018) focussed on grassland sites in central and northern Europe, Scotland and the United States. Using a multi-model approach, they concluded that grasslands can be exploited for GHG mitigation in milk and beef production, following C and N sequestrations, at least under certain circumstances, to offset GHG emissions. Sándor et al. (2020) was more limited in scope to European, Canadian as well as Indian grasslands and concluded that while GPP (gross primary production) and RECO (ecosystem respiration) tended to be underestimated, NEE (net ecosystem exchange) tended to be overestimated, so ensemble modelling was a useful approach to address model uncertainties.

## Model validation

While model evaluations have been conducted in other climates, those focussed on boreal regions are more scarce. The use of model validation methods varied between authors, some were extensive, for example He et al. (2020), though others used only a few methods over limited variables, for example (Chatskikh et al., 2005). The following main methods used in the various research papers reviewed describe validation results for studies carried out in boreal regions only. For BASGRA, three methods over three variables were used in two papers (Höglind et al., 2020; Korhonen et al., 2018). For DNDC, 12 methods over six variables (DNDC v.CAN was compared with DNDC for N leaching) were

obtained from two papers (Congreves et al., 2016; He et al., 2020), for FASSET four methods for one variable from a single paper (Chatskikh et al., 2005), and for IFSM five methods over four variables from one paper (Jégo et al., 2015). A single paper (Korhonen et al., 2018) tested forage production models STICS and CATIMO alongside BASGRA, giving four methods over three biomass variables (Tables S1, S5 and S6).

Equations and validation assessment tables are given in supplementary material. Our results are drawn from the few research articles found which include validation data for models in boreal climates. While limited in scope they highlight the scarcity of such studies. Some models, such as DNDC, BASGRA\_N and IFSM appear to perform within the range of 'fair' to 'good', though DNDC modelled soil chemistry poorly and BASGRA\_N also performed poorly at simulating biomass growth responses to N fertiliser. What can be seen from this table in addition to performance quality, is the lack of validation in a range of areas. This is partially a limitation of the models themselves as part of their design, for example, BASGRA\_N and CATIMO are both biomass focussed models, and DNDC and its extension model DNDC v.CAN are both nitrogen cycle-based models with gas exchange as a key area of focus (less so on production), though the models can certainly estimate these parameters, their accuracy remains uncertain.

## SYNTHESIS AND OUTLOOK

The objective of this review was to examine the use of models in grassland settings to assess potential for use in boreal grasslands. Overall, the use of relevant models in these regions was sparse, being mainly confined to a few models with variable results and degrees of applicability. We identified significant scope for model development and evaluation in boreal regions, particularly with respect to model intercomparison using datasets collected from boreal environments. We also identified a lack of information on (and to a lesser extent models equipped for) modelling land-use change, for example, conversion forestry, wetlands and so forth, to agricultural usages. Simulation of land-use change is important, as changes in land-use can lead to large effluxes of GHG emissions, partly due to carbon loss, for example, Laine et al. (2019). Although there is some literature documenting modelled changes from arable land use to grassland using DNDC and Century (Grant et al., 2004; Smith et al., 2001), and changes in grassland extent using ORCHIDEE\_GM (Chang et al., 2016) some of the models examined here only simulate elements of land use change (e.g., loss in soil carbon; APSIM, DNDC) while others may not be designed for substantive land use change at all. This may be because models tend to be designed for steady-state and prevailing conditions, rather than for abrupt transitions from one dominant land use (e.g., forestry) to another (e.g., cropping or pasture grazing).

The BASGRA\_N model was primarily designed for simulating biomass production, tillering dynamics and



pasture persistence but can also simulate GHG balance and is specifically designed with cold climate growth including winter conditions such as freezing soil and snow cover. However, the primary focus of BASGRA\_N has been on biomass production, especially its unique capacity to simulate root death, a key route for OM into the soil substrate, while a potential weakness in simulating fertiliser effects may indicate the need to calibrate the simulation components related to N cycling. The Day-Cent model has the potential to link atmospheric and soil models for boreal conditions, although it has not seen much use in these regions. It has performed well for modelling Canadian grassland in an ensemble modelling study.

The DNDC model is the most widely used and tested model in boreal regions and predicted N<sub>2</sub>O emissions, soil temperature and biomass production quite well. Simulations appear to be less reliable when estimating other key variables such as soil N leaching, and SOC and CO<sub>2</sub> fluxes, although the evaluation of the last two was not found in the papers reviewed. The DNDC v.CAN model has shown improvements in modelling both N leaching and NH<sub>3</sub> fluxes compared to the standard DNDC model and may be a preferable alternative for comparable soils in Finland.

The FASSET model, while potentially useful may be sensitive to soil structure and overestimate SOC and thus related N in undisturbed soils. Performance assessments have been limited to N<sub>2</sub>O emissions, where assessments were moderately satisfactory. The lack of studies in boreal areas for this model also needs to be addressed as far too little has been done to make a definitive assessment.

IFSM has been used successfully throughout North America to model multiple farm processes from runoff effects to biomass production and GHG emissions on a whole-farm basis and has also been used to predict the effects of climate change. IFSM provides a tool for similar studies within Finland, and although little work in this area has been carried out to date, existing results show fair performance in a comparable climate and so should warrant further exploration.

The ORCHIDEE\_GM model has shown good larger spatial-scale results including those for Scandinavia and so would be a good candidate model to investigate. ORCHIDEE\_GM has not been evaluated against site scale data in boreal regions, but was recently compared with detailed grassland production from multiple sites in Mongolia with good performance (Nandintsetseg et al., 2021).

Of the models examined here (Table 1), DNDC, IFSM and BASGRA\_N show the highest potential for simulating grassland and livestock productivity in boreal regions. The two models examined as part of the assessment of Korhonen et al. (2018), STICS and CATIMO also performed adequately, but being only part of a single study show only limited results. In terms of GHG simulations, DNDC again shows promise, and FASSET also has good potential although again data is limited. As yet no model provided CO<sub>2</sub> or CH<sub>4</sub> emission data validations, a key

area of interest in climate change mitigation, so further work is required here.

## Recommendations for future research

With a burgeoning global population, the need for both food security and increased agri-food production under a changing global climate is becoming more pressing. Intensification of global food supply and climate crisis adaptation must however occur without degrading natural capital, losing biodiversity, increasing GHG emissions or stimulating other adverse economic, environmental or social trade-offs. Process-based models have already proven as invaluable tools in a wide range of agro-ecological regions and at a range of scales, though such models remain underutilised in boreal regions. There is significant opportunity for future studies to identify models that have superior performance in the circumpolar regions (Roulet, 2000). This review has built a solid foundation for future research by identifying some of the more well-known models as candidates for further study as well as pointing out some of their weaknesses and areas of uncertainty. Model comparison studies have already been carried out in more general settings, such as the work of Ehrhardt et al. (2018) and Sándor et al. (2020), who have laid early foundations for this kind of research. It remains to expand this study into areas with harsher climates but great agricultural potential. Prime opportunities for future research include (1) evaluation of individual models using field data measured for grassland production, soil C and N cycling, and GHG emissions for multiple boreal environments, (2) evaluation of multiple models using the same datasets within boreal zones (i.e., a model intercomparison), (3) identification of sub-model processes underpinning superior model performance (e.g., soil C and N cycling, plant growth and senescence, simulation of soil water and reliability of holistic systems simulations, that is, the climate–soil–plant–livestock continuum) and (4) opportunities for parsimonisation to the proper level while keeping the model coherent.

## AUTHOR CONTRIBUTIONS

**Daniel Forster:** Conceptualization; data curation; methodology; project administration; visualization; writing – original draft; writing – review and editing. **Samuli Helama:** Data curation; writing – original draft; writing – review and editing. **Matthew Harrison:** Writing – review and editing. **Jinfeng Chang:** Data curation; writing – review and editing. **Elizabeth Pattey:** Writing – review and editing. **Perttu Virkajärvi:** Writing – review and editing. **Narasinha Shurpali:** Project administration; supervision; writing – review and editing.

## ACKNOWLEDGEMENTS

This study was supported by funding from the Ministry of Agriculture and Forestry Finland (Helsinki, FI) (Project: Clover for biogas, Project NC-GRASS: VN/28562/2020-MMM-2). We also acknowledge the support from the Academy of Finland funded ENSINK project (Decision number 334422).

## CONFLICTS OF INTEREST

The authors declare no conflicts of interest.

## DATA AVAILABILITY STATEMENT

Data sharing is not applicable to this article as no new data were created or analysed in this study.

## ORCID

Daniel Forster  <https://orcid.org/0000-0001-7514-7777>  
 Samuli Helama  <https://orcid.org/0000-0002-9777-3354>  
 Matthew T. Harrison  <https://orcid.org/0000-0001-7425-452X>  
 Clarence Alan Rotz  <https://orcid.org/0000-0001-6668-4319>  
 Jinfeng Chang  <https://orcid.org/0000-0003-4463-7778>  
 Phillippe Ciais  <https://orcid.org/0000-0001-8560-4943>  
 Elizabeth Pattey  <https://orcid.org/0000-0001-5082-5973>  
 Perttu Virkajärvi  <https://orcid.org/0000-0002-1954-9904>  
 Narasinha Shurpali  <https://orcid.org/0000-0003-1052-4396>

## REFERENCES

- Alcock, D. J., Harrison, M. T., Rawnsley, R. P., & Eckard, R. J. (2015). Can animal genetics and flock management be used to reduce greenhouse gas emissions but also maintain productivity of wool-producing enterprises? *Agricultural Systems*, 132, 25–34. <https://doi.org/10.1016/j.agsy.2014.06.007>
- Alemu, A. W., Ominski, K. H., Tenuta, M., Amiro, B. D., & Kebreab, E. (2016). Evaluation of greenhouse gas emissions from HOG manure application in a Canadian cow-calf production system using whole-farm models. *Animal Production Science*, 56(10), 1722–1737. <https://doi.org/10.1071/AN14994>
- Altdorff, D., Borchard, N., Young, E. H., Galagedara, L., Sorvali, J., Quideau, S., & Unc, A. (2021). Agriculture in boreal and Arctic regions requires an integrated global approach for research and policy. *Agronomy for Sustainable Development*, 41(2), 1–13. <https://doi.org/10.1007/s13593-021-00676-1>
- Ara, I., Turner, L., Harrison, M. T., Monjardino, M., deVoil, P., & Rodriguez, D. (2021). Application, adoption and opportunities for improving decision support systems in irrigated agriculture: A review. *Agricultural Water Management*, 257, 107161. <https://doi.org/10.1016/j.agwat.2021.107161>
- Bartlett, J., Rusch, G. M., Kyrkjeeide, M. O., Sandvik, H., & Nordin, J. (2020). Carbon storage in Norwegian ecosystems (revised edition). In *NINA Report, 1774b*, 1459–1473.
- Beaulne, J., Garneau, M., Magnan, G., & Boucher, É. (2021). Peat deposits store more carbon than trees in forested peatlands of the boreal biome. *Scientific Reports*, 11(1), 2657. <https://doi.org/10.1038/s41598-021-82004-x>
- Bilotto, F., Harrison, M. T., Migliorati, M. D. A., Christie, K. M., Rowlings, D. W., Grace, P. R., Smith, A. P., Rawnsley, R. P., Thorburn, P. J., & Eckard, R. J. (2021). Can seasonal soil N mineralisation trends be leveraged to enhance pasture growth? *Science of the Total Environment*, 772, 145031. <https://doi.org/10.1016/j.scitotenv.2021.145031>
- Boehm, M., Junkins, B., Desjardins, R., Kulshreshtha, S., & Lindwall, W. (2004). Sink potential of Canadian agricultural soils. *Climatic Change*, 65(3), 297–314. <https://doi.org/10.1023/B:CLIM.0000038205.09327.51>
- Bonesmo, H., & Bélanger, G. (2002). Timothy yield and nutritive value by the CATIMO Model: I. Growth and nitrogen. *Agronomy Journal*, 94(2), 337–345. <https://doi.org/10.2134/agronj2002.3370>
- Brisson, N., Gary, C., Justes, E., Roche, R., Mary, B., Ripoche, D., Zimmer, D., Sierra, J., Bertuzzi, P., Burger, P., Bussiere, F., Cabidoche, Y. M., Cellier, P., Debaeke, P., Gaudillere, J. P., Henault, C., Maraux, F., Seguin, B., & Sinoquet, H. (2003). An overview of the crop model STICS. *European Journal of Agronomy*, 18, 309–332. [https://doi.org/10.1016/S1161-0301\(02\)00110-7](https://doi.org/10.1016/S1161-0301(02)00110-7)
- Carlson, K. M., Gerber, J. S., Mueller, N. D., Herrero, M., MacDonald, G. K., Brauman, K. A., Havlik, P., O'Connell, C. S., Johnson, J. A., Saatchi, S., & West, P. C. (2017). Greenhouse gas emissions intensity of global croplands. *Nature Climate Change*, 7(1), 63–68. <https://doi.org/10.1038/nclimate3158>
- Chang, J., Ciais, P., Gasser, T., Smith, P., Herrero, M., Havlik, P., Obersteiner, M., Guenet, B., Goll, D. S., Li, W., Naipal, V., Peng, S., Qiu, C., Tian, H., Viovy, N., Yue, C., & Zhu, D. (2021). Climate warming from managed grasslands cancels the cooling effect of carbon sinks in sparsely grazed and natural grasslands. *Nature Communications*, 12(1), 1–10. <https://doi.org/10.1038/s41467-020-20406-7>
- Chang, J., Ciais, P., Viovy, N., Soussana, J.-F., Klumpp, K., & Sultan, B. (2017). Future productivity and phenology changes in European grasslands for different warming levels: Implications for grassland management and carbon balance. *Carbon balance and management*, 12(1), 11. <https://doi.org/10.1186/s13021-017-0079-8>
- Chang, J., Ciais, P., Viovy, N., Vuichard, N., Herrero, M., Havlik, P., Wang, X., Sultan, B., & Soussana, J.-F. (2016). Effect of climate change, CO<sub>2</sub> trends, nitrogen addition, and land-cover and management intensity changes on the carbon balance of European grasslands. *Global Change Biology*, 22(1), 338–350. <https://doi.org/10.1111/gcb.13050>
- Chang, J., Ciais, P., Viovy, N., Vuichard, N., Sultan, B., & Soussana, J.-F. (2015). The greenhouse gas balance of European grasslands. *Global Change Biology*, 21(10), 3748–3761. <https://doi.org/10.1111/gcb.12998>
- Chang, J., Viovy, N., Vuichard, N., Ciais, P., Campioli, M., Klumpp, K., Martin, R., Leip, A., & Soussana, J.-F. (2015). Modeled changes in potential grassland productivity and in grass-fed ruminant livestock density in Europe over 1961–2010. *PLOS ONE*, 10(5), e0127554. <https://doi.org/10.1371/journal.pone.0127554>
- Chang, J., Viovy, N., Vuichard, N., Ciais, P., Wang, T., Cozic, A., Lardy, R., Graux, A. I., Klumpp, K., Martin, R., & Soussana, J. F. (2013). Incorporating grassland management in ORCHIDEE: Model description and evaluation at 11 eddy-covariance sites in Europe. *Geoscientific Model Development*, 6(6), 2165–2181. <https://doi.org/10.5194/gmd-6-2165-2013>
- Chang-Fung-Martel, J., Harrison, M. T., Rawnsley, R., Smith, A. P., & Meinke, H. (2017). The impact of extreme climatic events on pasture-based dairy systems: A review. *Crop and Pasture Science*, 68(12), 1158. <https://doi.org/10.1071/CP16394>
- Chatskikh, D., Olesen, J. E., Berntsen, J., Regina, K., & Yamulki, S. (2005). Simulation of effects of soils, climate and management on N<sub>2</sub>O Emission from Grasslands. *Biogeochemistry*, 76(3), 395–419.
- Christie, K. M., Rawnsley, R. P., Harrison, M. T., & Eckard, R. J. (2014). Using a modelling approach to evaluate two options for improving animal nitrogen use efficiency and reducing nitrous oxide emissions on dairy farms in southern Australia. *Animal Production Science*, 54(12), 1960. <https://doi.org/10.1071/AN14436>
- Christie, K. M., Smith, A. P., Rawnsley, R. P., Harrison, M. T., & Eckard, R. J. (2018). Simulated seasonal responses of grazed dairy pastures to nitrogen fertilizer in SE Australia: Pasture production. *Agricultural Systems*, 166, 36–47. <https://doi.org/10.1016/j.agsy.2018.07.010>
- Christie, K. M., Smith, A. P., Rawnsley, R. P., Harrison, M. T., & Eckard, R. J. (2020). Simulated seasonal responses of grazed dairy pastures to nitrogen fertilizer in SE Australia: N loss and recovery. *Agricultural Systems*, 182, 102847. <https://doi.org/10.1016/j.agsy.2020.102847>
- Ciais, P., Reichstein, M., Viovy, N., Granier, A., Ogée, J., Allard, V., Aubinet, M., Buchmann, N., Bernhofer, C., Carrara, A., Chevallier, F., De Noblet, N., Friend, A. D., Friedlingstein, P., Grünwald, T., Heinesch, B., Keronen, P., Knohl, A., Krinner, G., & Valentini, R. (2005). Europe-wide reduction in primary productivity caused by the heat and drought in 2003. *Nature*, 437(7058), 529–533. <https://doi.org/10.1038/nature03972>
- Coleman, K., Jenkinson, D. S., Crocker, G. J., Grace, P. R., Klir, J., Körschens, M., Poulton, P. R., & Richter, D. D. (1997). Simulating trends in soil organic carbon in long-term

- experiments using RothC-26.3. *Geoderma*, 81(1–2), 29–44. [https://doi.org/10.1016/S0016-7061\(97\)00084-0](https://doi.org/10.1016/S0016-7061(97)00084-0)
- Congreves, K. A., Grant, B. B., Dutta, B., Smith, W. N., Chantigny, M. H., Rochette, P., & Desjardins, R. L. (2016). Predicting ammonia volatilization after field application of swine slurry: DNDC model development. *Agriculture, Ecosystems and Environment*, 219, 179–189. <https://doi.org/10.1016/j.agee.2015.10.028>
- Cordeiro, M. R. C., Rotz, A., Kroebe, R., Beauchemin, K. A., Hunt, D., Bittman, S., Koenig, K. M., & McKenzie, D. B. (2019). Prospects of forage production in northern regions under climate and land-use changes: A case-study of a dairy farm in Newfoundland, Canada. *Agronomy*, 9(31). <https://doi.org/10.3390/agronomy9010031>
- Del Grosso, S. J., Parton, W. J., Mosier, A. R., Hartman, M. D., Brenner, J., Ojima, D. S., & Schimel, D. (2001). Simulated interaction of carbon dynamics and nitrogen trace gas fluxes using the DAYCENT Model. In *Modeling carbon and nitrogen dynamics for soil management* (Issue May, pp. 303–332). <https://doi.org/10.1201/9781420032635.ch8>
- Desjardins, R. L., Smith, W., Grant, B., Campbell, C., Janzen, H., & Ritzek, R. (2005). Management strategies to sequester carbon in agricultural soils and to mitigate greenhouse gas emissions. *Climate Change*, 70(1), 283–297. <https://doi.org/10.1007/s10705-016-7-14>
- Despotovic, M., Nedic, V., Despotovic, D., & Cvetanovic, S. (2016). Evaluation of empirical models for predicting monthly mean horizontal diffuse solar radiation. *Renewable and Sustainable Energy Reviews*, 56, 246–260. <https://doi.org/10.1016/j.rser.2015.11.058>
- Duchemin, M., Jégo, G., & Morissette, R. (2019). Simulating switchgrass aboveground biomass and production costs in eastern Canada with the integrated farm system model. *Canadian Journal of Plant Science*, 99(6), 785–800. <https://doi.org/10.1139/cjps-2018-0331>
- Ehrhardt, F., Soussana, J., Bellocchi, G., Grace, P., McAuliffe, R., Recous, S., Sándor, R., Smith, P., Snow, V., deAntoni Migliorati, M., Basso, B., Bhatia, A., Brilli, L., Doltra, J., Dorich, C. D., Doro, L., Fitton, N., Giacomini, S. J., Grant, B., & Zhang, Q. (2018). Assessing uncertainties in crop and pasture ensemble model simulations of productivity and N<sub>2</sub>O emissions. *Global Change Biology*, 24(2), e603–e616. <https://doi.org/10.1111/gcb.13965>
- EOS. (2017). *DNDC. Scientific Basis and Processes* (Vol. 9.5; pp. 1–28). Institute for the Study of Earth, Oceans, and Space (EOS).
- Ergon, Å., Seddaiu, G., Korhonen, P., Virkajärvi, P., Bellocchi, G., Jørgensen, M., Østrem, L., Reheul, D., & Volaire, F. (2018). How can forage production in Nordic and Mediterranean Europe adapt to the challenges and opportunities arising from climate change? *European Journal of Agronomy*, 92(2017), 97–106. <https://doi.org/10.1016/j.eja.2017.09.016>
- Fan, J., McConkey, B. G., Liang, B. C., Angers, D. A., Janzen, H. H., Kröbel, R., Cerkowniak, D. D., & Smith, W. N. (2019). Increasing crop yields and root input make Canadian farmland a large carbon sink. *Geoderma*, 336, 49–58. <https://doi.org/10.1016/j.geoderma.2018.08.004>
- FAOSTAT. (2022). FAOSTAT Landuse. <https://www.fao.org/faostat/en/#data/RL/visualize>
- Farina, R., Sándor, R., Abdalla, M., Álvaro-Fuentes, J., Bechini, L., Bolinder, M. A., Brilli, L., Chenu, C., Clivot, H., De Antoni Migliorati, M., Di Bene, C., Dorich, C. D., Ehrhardt, F., Ferchaud, F., Fitton, N., Francaviglia, R., Franko, U., Giltrap, D. L., Grant, B. B., & Bellocchi, G. (2021). Ensemble modelling, uncertainty and robust predictions of organic carbon in long-term bare-fallow soils. *Global Change Biology*, 27(4), 904–928. <https://doi.org/10.1111/gcb.15441>
- Forster, D., Fraser, M. D., Rowe, R., & McNamara, N. P. (2021). Influence of liming and sward management on soil carbon storage by semi-improved upland grasslands. *Soil and Tillage Research*, 212, 105059. <https://doi.org/10.1016/j.still.2021.105059>
- Glasgow climate change conference—October–November. (2021). United Nations Climate Change. <https://unfccc.int/documents/310475>
- Grant, B., Smith, W. N., Desjardins, R., Lemke, R., & Li, C. (2004). Estimated N<sub>2</sub>O And CO<sub>2</sub> emissions as influenced by agricultural practices in Canada. *Climate Change*, 65, 315–332. <https://doi.org/10.1023/B:CLIM.0000038226.60317.35>
- Guest, G., Kröbel, R., Grant, B., Smith, W., Sansoulet, J., Pattey, E., Desjardins, R., Jégo, G., Tremblay, N., & Tremblay, G. (2017). Model comparison of soil processes in eastern Canada using DayCent, DNDC and STICS. *Nutrient Cycling in Agroecosystems*, 109(3), 211–232. <https://doi.org/10.1007/s10705-017-9880-8>
- Harrison, M. T., Christie, K. M., Rawnsley, R. P., & Eckard, R. J. (2014). Modelling pasture management and livestock genotype interventions to improve whole-farm productivity and reduce greenhouse gas emissions intensities. *Animal Production Science*, 54(12), 2018. <https://doi.org/10.1071/AN14421>
- Harrison, M. T., Cullen, B. R., & Armstrong, D. (2017). Management options for dairy farms under climate change: Effects of intensification, adaptation and simplification on pastures, milk production and profitability. *Agricultural Systems*, 155, 19–32. <https://doi.org/10.1016/j.agsy.2017.04.003>
- Harrison, M. T., Cullen, B. R., Mayberry, D. E., Cowie, A. L., Bilotto, F., Badgery, W. B., Liu, K., Davison, T., Christie, K. M., Muleke, A., & Eckard, R. J. (2021). Carbon myopia: The urgent need for integrated social, economic and environmental action in the livestock sector. *Global Change Biology*, 27(22), 5726–5761. <https://doi.org/10.1111/gcb.15816>
- Harrison, M. T., Cullen, B. R., Tomkins, N. W., McSweeney, C., Cohn, P., & Eckard, R. J. (2016). The concordance between greenhouse gas emissions, livestock production and profitability of extensive beef farming systems. *Animal Production Science*, 56(3), 370. <https://doi.org/10.1071/AN15515>
- Harrison, M. T., Evans, J. R., Dove, H., & Moore, A. D. (2011). Recovery dynamics of rainfed winter wheat after livestock grazing 2. Light interception, radiation-use efficiency and dry-matter partitioning. *Crop and Pasture Science*, 62(11), 960. <https://doi.org/10.1071/CP11235>
- Harrison, M. T., Evans, J. R., & Moore, A. D. (2012a). Using a mathematical framework to examine physiological changes in winter wheat after livestock grazing. *Field Crops Research*, 136, 116–126. <https://doi.org/10.1016/j.fcr.2012.06.015>
- Harrison, M. T., Evans, J. R., & Moore, A. D. (2012b). Using a mathematical framework to examine physiological changes in winter wheat after livestock grazing. *Field Crops Research*, 136, 127–137. <https://doi.org/10.1016/j.fcr.2012.06.014>
- Harrison, M. T., Jackson, T., Cullen, B. R., Rawnsley, R. P., Ho, C., Cummins, L., & Eckard, R. J. (2014). Increasing ewe genetic fecundity improves whole-farm production and reduces greenhouse gas emissions intensities. *Agricultural Systems*, 131, 23–33. <https://doi.org/10.1016/j.agsy.2014.07.008>
- Harrison, M. T., Roggero, P. P., & Zavattaro, L. (2019). Simple, efficient and robust techniques for automatic multi-objective function parameterisation: Case studies of local and global optimisation using APSIM. *Environmental Modelling & Software*, 117, 109–133. <https://doi.org/10.1016/j.envsoft.2019.03.010>
- He, W., Dutta, B., Grant, B. B., Chantigny, M. H., Hunt, D., Bittman, S., Tenuta, M., Worth, D., VanderZaag, A., Desjardins, R. L., & Smith, W. N. (2020). Assessing the effects of manure application rate and timing on nitrous oxide emissions from managed grasslands under contrasting climate in Canada. *Science of the Total Environment*, 716, 135374. <https://doi.org/10.1016/j.scitotenv.2019.135374>
- Heikkinen, J., Ketoja, E., Nuutinen, V., & Regina, K. (2013). Declining trend of carbon in Finnish cropland soils in 1974–2009. *Global Change Biology*, 19(5), 1456–1469. <https://doi.org/10.1111/gcb.12137>
- Heimsch, L., Lohila, A., Tuovinen, J. P., Vekuri, H., Heinonsalo, J., Nevalainen, O., Korkiakoski, M., Liski, J., Laurila, T., & Kulmala, L. (2021). Carbon dioxide fluxes and carbon balance of an agricultural grassland in southern Finland. *Biogeosciences*, 18(11), 3467–3483. <https://doi.org/10.5194/bg-18-3467-2021>
- Helgadóttir, A., Frankow-Lindberg, B. E., Seppänen, M. M., Søgaard, K., & Østrem, L. (2014). European grasslands



- overview: Nordic region. *Grassland Science in Europe*, Vol. 19 - EGF at 50: The Future of European Grasslands, 19, 15–25.
- Ho, C. K. M., Jackson, T., Harrison, M. T., & Eckard, R. J. (2014). Increasing ewe genetic fecundity improves whole-farm production and reduces greenhouse gas emissions intensities: 2. Economic performance. *Animal Production Science*, 54(9), 1248. <https://doi.org/10.1071/AN14309>
- Höglind, M., Cameron, D., Persson, T., Huang, X., & vanOijen, M. (2020). BASGRA\_N: A model for grassland productivity, quality and greenhouse gas balance. *Ecological Modelling*, 417(January), 108925. <https://doi.org/10.1016/j.ecolmodel.2019.108925>
- Höglind, M., Thorsen, S. M., & Semenov, M. A. (2013). Assessing uncertainties in impact of climate change on grass production in Northern Europe using ensembles of global climate models. *Agricultural and Forest Meteorology*, 170, 103–113. <https://doi.org/10.1016/j.agrformet.2012.02.010>
- Höglind, M., Van Oijen, M., Cameron, D., & Persson, T. (2016). Process-based simulation of growth and overwintering of grassland using the BASGRA model. *Ecological Modelling*, 335, 1–15. <https://doi.org/10.1016/j.ecolmodel.2016.04.024>
- Hugelius, G., Loisel, J., Chadburn, S., Jackson, R. B., Jones, M., MacDonald, G., Marushchak, M., Olefeldt, D., Packalen, M., Siewert, M. B., Treat, C., Turetsky, M., Voigt, C., & Yu, Z. (2020). Large stocks of peatland carbon and nitrogen are vulnerable to permafrost thaw. *Proceedings of the National Academy of Sciences*, 117(34), 20438–20446. <https://doi.org/10.1073/pnas.1916387117>
- IPCC. (2021). *Summary for Policymakers*. In V. Masson-Delmotte, P. Zhai, A. Pirani, S. L. Connors, C. Péan, S. Berger, N. Caud, Y. Chen, L. Goldfarb, M. I. Gomis, M. Huang, K. Leitzell, E. Lonnoy, J. B. R. Matthews, T. K. Maycock, T. Waterfield, O. Yelekçi, R. Yu, & B. Zhou (Eds.), *Climate change 2021: The physical science basis. Contribution of Working Group I to the Sixth Assessment Report of the Intergovernmental Panel on Climate Change* (pp. 3–32). Cambridge University Press.
- Jacobsen, B. H., Boye, C., Petersen, B. M., Bernsten, J., Sørensen, C. G., Søgaard, H. T., & Hansen, J. P. (1998). An Integrated Economic and Environmental Farm Simulation Model (FASSET). Danish Institute of Agricultural and Fisheries Economics.
- Jégo, G., Chantigny, M., Pattey, E., Bélanger, G., Rochette, P., Vanasse, A., & Goyer, C. (2014). Improved snow-cover model for multi-annual simulations with the STICS crop model under cold, humid continental climates. *Agricultural and Forest Meteorology*, 195–196, 38–51. <https://doi.org/10.1016/j.agrformet.2014.05.002>
- Jégo, G., Rotz, C. A., Bélanger, G., Tremblay, G. F., Charbonneau, É., & Pellerin, D. (2015). Simulating forage crop production in a northern climate with the Integrated Farm System Model. *Canadian Journal of Plant Science*, 95(4), 745–757. <https://doi.org/10.4141/cjps-2014-375>
- Jones, J. W., Antle, J. M., Basso, B., Boote, K. J., Conant, R. T., Foster, I., Godfray, H. C. J., Herrero, M., Howitt, R. E., Janssen, S., Keating, B. A., Munoz-Carpena, R., Porter, C. H., Rosenzweig, C., & Wheeler, T. R. (2016). Brief history of agricultural systems modeling. *Agricultural Systems*, 155, 240–254. <https://doi.org/10.1016/j.agry.2016.05.014>
- Jose, V. S., Sejian, V., Bagath, M., Ratnakaran, A. P., Lees, A. M., Al-Hosni, Y. A. S., Sullivan, M., Bhatta, R., & Gaughan, J. B. (2016). Modeling of greenhouse gas emission from livestock. *Frontiers in Environmental Science*, 4(APR), 1–10. <https://doi.org/10.3389/fenvs.2016.00027>
- Kätterer, T., Bolinder, M. A., Andrén, O., Kirchmann, H., & Menichetti, L. (2011). Roots contribute more to refractory soil organic matter than above-ground crop residues, as revealed by a long-term field experiment. *Agriculture, Ecosystems & Environment*, 141(1–2), 184–192. <https://doi.org/10.1016/j.agee.2011.02.029>
- Khalil, M. I., Fornara, D. A., & Osborne, B. (2020). Simulation and validation of long-term changes in soil organic carbon under permanent grassland using the DNDC model. *Geoderma*, 361(2019), 114014. <https://doi.org/10.1016/j.geoderma.2019.114014>
- Kim, D., Stoddart, N., Rotz, C. A., Veltman, K., Chase, L., Cooper, J., Ingraham, P., Izaurralde, R. C., Jones, C. D., Gaillard, R., Aguirre-Villegas, H. A., Larson, R. A., Ruark, M., Salas, W., Jolliet, O., & Thoma, G. J. (2019). Analysis of beneficial management practices to mitigate environmental impacts in dairy production systems around the Great Lakes. *Agricultural Systems*, 176. <https://doi.org/10.1016/j.agry.2019.102660>
- Kim, M. K., & Henry, H. (2013). Net ecosystem CO<sub>2</sub> exchange and plant biomass responses to warming and N addition in a grass-dominated system during two years of net CO<sub>2</sub> efflux. *Plant and Soil*, 371(1–2), 409–421. <https://doi.org/10.1007/s11104-013-1705-1>
- Korhonen, P., Palosuo, T., Persson, T., Höglind, M., Jégo, G., Van Oijen, M., Gustavsson, A. M., Bélanger, G., & Virkajärvi, P. (2018). Modelling grass yields in northern climates—A comparison of three growth models for timothy. *Field Crops Research*, 224, 37–47. <https://doi.org/10.1016/j.fcr.2018.04.014>
- Krinner, G., Viovy, N., deNoblet-Ducoudré, N., Ogée, J., Polcher, J., Friedlingstein, P., Ciais, P., Sitch, S., & Prentice, I. C. (2005). A dynamic global vegetation model for studies of the coupled atmosphere-biosphere system. *Global Biogeochemical Cycles*, 19(1), 1–33. <https://doi.org/10.1029/2003GB002199>
- Laine, A. M., Mehtätalo, L., Tolvanen, A., Frohling, S., & Tuittila, E.-S. (2019). Impacts of drainage, restoration and warming on boreal wetland greenhouse gas fluxes. *Science of the Total Environment*, 647, 169–181. <https://doi.org/10.1016/j.scitotenv.2018.07.390>
- Leblans, N. I. W., Sigurdsson, B. D., Aerts, R., Vicca, S., Magnússon, B., & Janssens, I. A. (2017). Icelandic grasslands as long-term C sinks under elevated organic N inputs. *Biogeochemistry*, 134(3), 279–299. <https://doi.org/10.1007/s10533-017-0362-5>
- Levy, P. E., Mobbs, D. C., Jones, S. K., Milne, R., Campbell, C., & Sutton, M. A. (2007). Simulation of fluxes of greenhouse gases from European grasslands using the DNDC model. *Agriculture, Ecosystems and Environment*, 121(1–2), 186–192. <https://doi.org/10.1016/j.agee.2006.12.019>
- Lind, S. E., Virkajärvi, P., Hyvönen, N. P., Maljanen, M., Kivimäenpää, M., Jokinen, S., Antikainen, S., Latva, M., Rätty, M., Martikainen, P. J., & Shurpali, N. J. (2020). Carbon dioxide and methane exchange of a perennial grassland on a boreal mineral soil. *Boreal Environment Research*, 25, 1–17.
- Lohila, A. (2004). Annual CO<sub>2</sub> exchange of a peat field growing spring barley or perennial forage grass. *Journal of Geophysical Research*, 109(D18), D18116. <https://doi.org/10.1029/2004JD004715>
- Lugato, E., Bampa, F., Panagos, P., & Montanarella, L. (2015). Potential carbon sequestration of European arable soils estimated by modelling a comprehensive set of management practices. *Global Change Biology*, 20, 3557–3567. <https://doi.org/10.1111/gcb.12551>
- Lundblad, M. (2015). *Land use on organic soils in Sweden. An overview on agriculture, forest lands and land use changes on organic soils* (Vol. 199). Swedish University of Agricultural Sciences.
- Maljanen, M., Sigurdsson, B. D., Guðmundsson, J., Óskarsson, H., Huttunen, J. T., & Martikainen, P. J. (2010). Greenhouse gas balances of managed peatlands in the Nordic countries—present knowledge and gaps. *Biogeosciences*, 7(9), 2711–2738. <https://doi.org/10.5194/bg-7-2711-2010>
- Mardian, J., Berg, A., & Daneshfar, B. (2021). Evaluating the temporal accuracy of grassland to cropland change detection using multitemporal image analysis. *Remote Sensing of Environment*, 255, 112292. <https://doi.org/10.1016/j.rse.2021.112292>
- Metzger, M. J., Bunce, R. G. H., Jongman, R. H. G., Múcher, C. A., & Watkins, J. W. (2005). A climatic stratification of the environment of Europe. *Global Ecology and Biogeography*, 14(6), 549–563. <https://doi.org/10.1111/j.1466-822X.2005.00190.x>
- Minasny, B., Malone, B. P., McBratney, A. B., Angers, D. A., Arrouays, D., Chambers, A., Chaplot, V., Chen, Z. S., Cheng, K., Das, B. S., Field, D. J., Gimona, A., Hedley, C. B., Hong, S. Y., Mandal, B., Marchant, B. P., Martin, M., McConkey, B. G., Mulder, V. L., & Winowiecki, L. (2017). Soil carbon 4 per mille. *Geoderma*, 292, 59–86. <https://doi.org/10.1016/j.geoderma.2017.01.002>

- Miura, M., Hill, P. W., & Jones, D. L. (2020). Impact of a single freeze-thaw and dry-wet event on soil solutes and microbial metabolites. *Applied Soil Ecology*, 153, 103636. <https://doi.org/10.1016/j.apsoil.2020.103636>
- Moriasi, D. N., Gitau, M. W., Pai, N., & Daggupati, P. (2015). Hydrologic and water quality models: Performance measures and evaluation criteria. *Transactions of the ASABE*, 58(6), 1763–1785. <https://doi.org/10.13031/trans.58.10715>
- Myllis, M., Lilja, H., & Regina, K. (2012). The Area of Cultivated Organic Soils in Finland According To GIS Datasets. *14th International Peat Congress: Extended Abstract No. 241*, 5.
- Nandintsetseg, B., Boldgiv, B., Chang, J., Ciais, P., Davaanyam, E., Batbold, A., Bat-Oyun, T., & Stenseth, N. C. (2021). Risk and vulnerability of Mongolian grasslands under climate change. *Environmental Research Letters*, 16(3), 034035. <https://doi.org/10.1088/1748-9326/abdb5b>
- Niemi, J., & Minna, V. (2019). *Agriculture and food sector in Finland 2019* (J. Niemi & V. Minna, Eds.). Natural Resources Institute Finland (Luke).
- Official Statistics of Finland. (2017). *Greenhouse Gases*. Statistics Finland. [https://www.stat.fi/til/khki/2017/01/khki\\_2017\\_01\\_2019-01-15\\_tie\\_001\\_en.html](https://www.stat.fi/til/khki/2017/01/khki_2017_01_2019-01-15_tie_001_en.html)
- Olesen, J. E., Petersen, B. M., Bernsten, J., Hansen, S., Jamieson, P. D., & Thomsen, A. G. (2002). Comparison of methods for simulating effects of nitrogen on green area index and dry matter growth in winter wheat. *Field Crops Research*, 74(2–3), 131–149. [https://doi.org/10.1016/S0378-4290\(01\)00204-0](https://doi.org/10.1016/S0378-4290(01)00204-0)
- Paris Agreement to the United Nations Framework Convention on Climate Change, 1 (2015).
- Parton, W. J., Hartman, M., Ojima, D., & Schimel, D. (1998). DAYCENT and its land surface submodel: Description and testing. *Global and Planetary Change*, 19(1–4), 35–48. [https://doi.org/10.1016/S0921-8181\(98\)00040-X](https://doi.org/10.1016/S0921-8181(98)00040-X)
- Persson, T., Höglind, M., Van Oijen, M., Korhonen, P., Palosuo, T., Jégo, G., Virkajärvi, P., Bélanger, G., & Gustavsson, A. M. (2019). Simulation of timothy nutritive value: A comparison of three process-based models. *Field Crops Research*, 231(2018), 81–92. <https://doi.org/10.1016/j.fcr.2018.11.008>
- Phelan, D. C., Harrison, M. T., Kemmerer, E. P., & Parsons, D. (2015). Management opportunities for boosting productivity of cool-temperate dairy farms under climate change. *Agricultural Systems*, 138, 46–54. <https://doi.org/10.1016/j.agry.2015.05.005>
- Piao, S., Friedlingstein, P., Ciais, P., De Noblet-Ducoudré, N., Labat, D., & Zaehle, S. (2007). Changes in climate and land use have a larger direct impact than rising CO<sub>2</sub> on global river runoff trends. *Proceedings of the National Academy of Sciences of the United States of America*, 104(39), 15242–15247. <https://doi.org/10.1073/pnas.0707213104>
- Poeplau, C., Aronsson, H., Myrbeck, Å., & Kätterer, T. (2015). Effect of perennial ryegrass cover crop on soil organic carbon stocks in southern Sweden. *Geoderma Regional*, 4, 126–133. <https://doi.org/10.1016/j.geoder.2015.01.004>
- Qiu, C., Ciais, P., Zhu, D., Guenet, B., Peng, S., Petrescu, A. M. R., Lauerwald, R., Makowski, D., Gallego-Sala, A. V., Charman, D. J., & Brewer, S. C. (2021). Large historical carbon emissions from cultivated northern peatlands. *Science Advances*, 7(23), eabf1332. <https://doi.org/10.1126/sciadv.abf1332>
- Rawnsley, R., Dynes, R. A., Christie, K. M., Harrison, M. T., Doran-Browne, N. A., Vibart, R., & Eckard, R. (2018). A review of whole farm-system analysis in evaluating greenhouse-gas mitigation strategies from livestock production systems. *Animal Production Science*, 58(6), 980. <https://doi.org/10.1071/AN15632>
- Rotz, A. C., Corson, M. S., Chianese, D. S., Montes, F., Hafner, S. D., Bonifacio, H. F., & Coiner, C. U. (2018). The integrated farm system model. In *Pasture systems and watershed research unit agricultural research service*. (Vol. 4.4; p. 249). USDA.
- Roulet, N. T. (2000). Peatlands, carbon storage, greenhouse gases, and the Kyoto Protocol: Prospects and significance for Canada. *Wetlands*, 20(4), 605–615. [https://doi.org/10.1672/0277-5212\(2000\)020\[0605:PCSGGA\]2.0.CO;2](https://doi.org/10.1672/0277-5212(2000)020[0605:PCSGGA]2.0.CO;2)
- Sándor, R., Ehrhardt, F., Basso, B., Bellocchi, G., Bhatia, A., Brilli, L., Migliorati, M. D., Doltra, J., Dorich, C., Doro, L., Fitton, N., Giacomini, S. J., Grace, P., Grant, B., Harrison, M. T., Jones, S., Kirschbaum, M. U. F., Klumpp, K., Laville, P., & Soussana, J. F. (2016). C and N models Intercomparison—benchmark and ensemble model estimates for grassland production. *Advances in Animal Biosciences*, 7(3), 245–247. <https://doi.org/10.1017/S2040470016000297>
- Sándor, R., Ehrhardt, F., Brilli, L., Carozzi, M., Recous, S., Smith, P., Snow, V., Soussana, J. F., Dorich, C. D., Fuchs, K., Fitton, N., Gongadze, K., Klumpp, K., Liebig, M., Martin, R., Merbold, L., Newton, P. C. D., Rees, R. M., Rolinski, S., & Bellocchi, G. (2018). The use of biogeochemical models to evaluate mitigation of greenhouse gas emissions from managed grasslands. *Science of the Total Environment*, 642(February), 292–306. <https://doi.org/10.1016/j.scitotenv.2018.06.020>
- Sándor, R., Ehrhardt, F., Grace, P., Recous, S., Smith, P., Snow, V., Soussana, J. F., Basso, B., Bhatia, A., Brilli, L., Doltra, J., Dorich, C. D., Doro, L., Fitton, N., Grant, B., Harrison, M. T., Kirschbaum, M. U. F., Klumpp, K., Laville, P., & Bellocchi, G. (2020). Ensemble modelling of carbon fluxes in grasslands and croplands. *Field Crops Research*, 252(March), 107791. <https://doi.org/10.1016/j.fcr.2020.107791>
- Sansoulet, J., Pattey, E., Kröbel, R., Grant, B., Smith, W., Jégo, G., Desjardins, R. L., Tremblay, N., & Tremblay, G. (2014). Comparing the performance of the STICS, DNDC, and DayCent models for predicting N uptake and biomass of spring wheat in Eastern Canada. *Field Crops Research*, 156, 135–150. <https://doi.org/10.1016/j.fcr.2013.11.010>
- Shcherbak, I., Millar, N., & Robertson, G. P. (2014). Global meta-analysis of the nonlinear response of soil nitrous oxide (N<sub>2</sub>O) emissions to fertilizer nitrogen. *Proceedings of the National Academy of Sciences*, 111(25), 9199–9204. <https://doi.org/10.1073/pnas.1322434111>
- Smith, W. N., Desjardins, R. L., & Grant, B. (2001). Estimated changes in soil carbon associated with agricultural practices in Canada. *Canadian Journal of Soil Science*, 81(2), 221–227. <https://doi.org/10.4141/S00-033>
- Soussana, J. F., Tallec, T., & Blanfort, V. (2010). Mitigating the greenhouse gas balance of ruminant production systems through carbon sequestration in grasslands. *International Conference Livestock and Global Climate Change*, 4(3), 334–350. <https://doi.org/10.1017/S1751731109990784>
- Taylor, C. A., Harrison, M. T., Telfer, M., & Eckard, R. (2016). Modelled greenhouse gas emissions from beef cattle grazing irrigated leucaena in northern Australia. *Animal Production Science*, 56(3), 594. <https://doi.org/10.1071/AN15575>
- Thivierge, M. N., Jégo, G., Bélanger, G., Chantigny, M. H., Rotz, C. A., Charbonneau, É., Baron, V. S., & Qian, B. (2017). Projected impact of future climate conditions on the agronomic and environmental performance of Canadian dairy farms. *Agricultural Systems*, 157(June), 241–257. <https://doi.org/10.1016/j.agry.2017.07.003>
- Twine, R. (2021). Emissions from animal agriculture—16.5% is the new minimum figure. *Sustainability*, 13(11), 6276. <https://doi.org/10.3390/su13116276>
- Unc, A., Altdorff, D., Abakumov, E., Adl, S., Baldursson, S., Bechtold, M., Cattani, D. J., Firbank, L. G., Grand, S., Guðjónsdóttir, M., Kallenbach, C., Kadir, A. J., Li, P., McKenzie, D. B., Misra, D., Nagano, H., Neher, D. A., Niemi, J., Oelbermann, M., & Borchard, N. (2021). Expansion of agriculture in Northern cold-climate regions: A cross-sectoral perspective on opportunities and challenges. *Frontiers in Sustainable Food Systems*, 5, 663448. <https://doi.org/10.3389/fsufs.2021.663448>
- Whitehead, D. (2020). Management of grazed landscapes to increase soil carbon stocks in temperate, Dryland Grasslands. *Frontiers in Sustainable Food Systems*, 4(October), 1–7. <https://doi.org/10.3389/fsufs.2020.585913>
- Wilberforce, T., Olabi, A. G., Sayed, E. T., Elsaid, K., & Abdelkareem, M. A. (2021). Progress in carbon capture technologies. *Science of the Total Environment*, 761, 143203. <https://doi.org/10.1016/j.scitotenv.2020.143203>
- deWit, H. A., Austnes, K., Hylen, G., & Dalsgaard, L. (2015). A carbon balance of Norway: Terrestrial and aquatic carbon fluxes. *Biogeochemistry*, 123(1–2), 147–173. <https://doi.org/10.1007/s10533-014-0060-5>

Xu, X., Sharma, P., Shu, S., Lin, T.-S., Ciais, P., Tubiello, F. N., Smith, P., Campbell, N., & Jain, A. K. (2021). Global greenhouse gas emissions from animal-based foods are twice those of plant-based foods. *Nature Food*, 2(9), 724–732. <https://doi.org/10.1038/s43016-021-00358-x>

## SUPPORTING INFORMATION

Additional supporting information can be found online in the Supporting Information section at the end of this article.

**How to cite this article:** Forster, D., Helama, S., Harrison, M. T., Rotz, C. A., Chang, J., Ciais, P., Pattey, E., Virkajärvi, P., & Shurpali, N. (2022). Use, calibration and verification of agroecological models for boreal environments: A review. *Grassland Research*, 1(1), 14–30. <https://doi.org/10.1002/glr2.12010>